Ph.D. DISSERTATION

FACULTY OF WOOD ENGINEERING AND CREATIVE INDUSTRIES

UNIVERSITY OF SOPRON

by

CHENAR A. TAHIR

SOPRON, HUNGARY

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Air-cathode ML-MFC reactor configuration using wood as container and separator to prevent deterioration and biofilm formation on cathode surface

A DISSERTATION

submitted in partial fulfilment of the requirements for the award of the degree of **DOCTOR OF PHILOSOPHY**

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under the supervision of **PROF. DR. LEVENTE CSÓKA DR. ZOLTÁN PÁSZTORY**

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Dissertation for doctoral (PhD) degree University of Sopron József Cziráki Doctoral School of Wood Sciences and Technologies, P4 Fiber and nanotechnology programme

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DECLARATION

I undersigned **Chenar Ali Tahir** by signing this declaration declare that my PhD thesis entitled **"Air-cathode ML-MFC reactor configuration using wood as container and separator to prevent deterioration and biofilm formation on cathode surface"** was my own work; during the dissertation I complied with the regulations of Act LXXVI of 1999 on Copyright and the rules of the doctoral dissertation prescribed by the Cziráki József Doctoral School, especially regarding references and citations[.](#page-3-0)¹

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This work is dedicated to my parents, brothers and sisters, for always being there and supporting me...

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Chenar A. Tahir

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ABSTRACT

In this study, we introduce a novel application of wood as the construction material for Microbial Fuel Cell (MFC) containers, developing what we term the Wooden Membrane-less Microbial Fuel Cell (WML-MFC). This innovative approach leverages the natural properties of wood to enhance the sustainability and cost-effectiveness of MFCs. Employing three different types of wood—pine, oak, and black locust—the WML-MFC design incorporates a carbon felt anode inside the wooden container and a carbon cloth cathode wrapped externally. This configuration not only protects the cathode from biofouling but also utilizes the inherent moisture management capabilities of wood to maintain operational stability. The performance of these wooden MFCs was assessed in terms of electricity generation and water treatment efficacy. Pine and oak containers achieved maximum power densities (MPD) of 35 mW/m² and 4 mW/m², respectively, with corresponding maximum open-circuit voltages of 551 mV and 269 mV. Black locust showed the least effective bioelectricity generation. COD removal efficiency was observed between 18% and 48% for pine and 3% to 39% for oak over hydraulic retention times of 24-48 hours. Notable water loss due to moisture diffusion was recorded at 20%/d in pine and 6%/d in oak. Durability assessments through DMA and SEM analyses confirmed the suitability of wood as a container material, emphasizing the dual environmental and economic benefits of this WML-MFC design. Notably, 2 mm and 3 mm thick Scots pine and oak containers exhibited lower electricity production, reduced water treatment capacity, higher water loss with thinner walls, and increased voltage fluctuations compared to the 4 mm thick containers.

Briefly, This study demonstrates that the wooden membrane-less microbial fuel cell (WML-MFC) system offers a cost-effective, environmentally friendly, and sustainable approach. It effectively safeguards the cathode from biofilm formation, thus preventing its deterioration and showcasing its potential as an eco-friendly energy generation and water treatment technology.

Keywords: Membrane-less microbial fuel cell, cathode deterioration, biofilm, wooden container

CHAPTER I-INTRODUCTION AND LITERATURE REVIEW

1. INTRODUCTION

1.1. Chapter synopsis

This chapter serves as a concise introduction and literature review focusing on various types of microbial fuel cells (MFCs). These MFC categories include MFCs with membranes, membrane-less MFCs (ML-MFCs), air-cathode MFCs, aqueous-cathode MFCs, and their variations with single or dual chamber designs. The subsequent sections delve deeper into membrane-less MFCs, examining recent research and developments using both traditional and contemporary testing techniques. Emphasis is placed on the applications in water treatment and electricity production. Additionally, this chapter provides an overview of the challenges and limitations within MFC technology, laying the foundation for the problem statement and study objectives. In conclusion, we offer an outline for the entire dissertation.

1.2. Problem statement and scope of the study

In recent decades, global concerns surrounding sustainability have driven the scientific community to seek innovative solutions for environmental issues (Nastro, 2014). Simultaneously, the depletion of crude oil resources has led to an increased focus on alternative energy sources. In this context, microbial fuel cells (MFCs) have gained international attention as a sustainable technology capable of generating electricity from organic matter in wastewater, while also aiding in wastewater treatment and environmental remediation (Feng et al., 2008). MFC technology holds the potential to contribute significantly to future energy needs (Logan, 2010). Moreover, it is costeffective, self-sustaining, and, in some cases, does not require substantial capital investment.

Among MFC researchers, membrane-less microbial fuel cells (ML-MFCs) have shown superior performance compared to membrane-based MFCs. Air-cathode MFCs are preferred over aqueous-cathode counterparts due to their independence from an external air supply. However, cathode deterioration remains a common issue across all membrane-less air-cathode designs, primarily caused by biofilm formation on the cathode's surface due to contact with the anolyte. This problem leads to a gradual decline in cell efficiency over time, making MFCs economically unviable. Selecting appropriate designs and materials for constructing MFC units becomes crucial. Conventional ML-MFC unit construction is often labor-intensive, time-consuming, environmentally unfriendly, and costly. To address

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these drawbacks, this study explores eco-friendly and readily available materials, such as wood, in ML-MFC unit construction. Wood exhibits numerous appealing properties, including biodegradability, porosity, accessibility, resistance to microbial attacks, and long lifespan.

1.3. Energy harvesting techniques

The world is experiencing a surge in electrical energy generation from primary environmental energy sources, including chemical (Rabaey and Verstraete, 2005), thermal sources such as geothermal and solar thermal (Kodama, 2003), vibrational energy (Ambrozkiewicz and Wolszczak, 2020), radio frequency (ElAnzeery et al., 2012), and mechanical energy sources such as water motion (Helseth and Guo, 2016). Additionally, modern technological devices consume less energy, especially small-sized devices and sensors. This makes microbial fuel cells a viable option for new biosensor applications and small devices that operate on battery power (Ivars-Barceló et al., 2018).

1.4. Background on Microbial fuel cell

The concept of microbial fuel cells (MFCs) was initially proposed by Michael Cressé Potter in 1912, utilizing Escherichia coli and Saccharomyces with platinum electrodes. However, low power output limited its early adoption (Yang et al., 2011). In the early 1990s, renewed interest in MFCs arose as advancements led to improved efficiency (Rahimnejad et al., 2015). In 1999, Kim et al. reported the first mediator-less MFC (Logan, 2008), paving the way for further developments and increased power density.

MFCs function by converting chemical energy from organic matter in wastewater directly into electrical energy with the assistance of electrogenic bacteria (Min and Logan, 2004) -Various microorganisms are capable of producing electrons through the metabolism of organic matter - (Cheng et al., 2006; Liu et al., 2004). An MFC typically consists of three components: the electrode system, a microorganism culture (either anaerobic or aerobic) fed with organic materials, and a container with a proton exchange membrane (PEM). MFCs come in different configurations, primarily single-chamber or dual-chamber designs (Khan et al., 2018), with most single-chamber MFCs incorporating a PEM to separate the anode and cathode compartments.

1.5. Background on membrane-less microbial fuel cell (ML-MFC)

Membranes in traditional MFCs introduce significant internal resistance, negatively impacting electrochemical performance and rendering them economically unviable (Logan, 2010). Efforts have been directed towards achieving higher output without membranes, leading to the emergence of membrane-less microbial fuel cells (ML-MFCs). These ML-MFCs offer several advantages, including energy selfsufficiency, operation at ambient temperatures, environmental friendliness by avoiding toxic by-products, reduced manufacturing costs, and extended operational lifespans (Fig. 1.2) (Yang et al., 2009).

ML-MFCs address the membrane-related disadvantages of traditional MFCs, where membranes impede functionality by increasing internal resistance (Jang et al., 2004). These devices hold the key to achieving higher efficiency by eliminating the membrane, thereby enhancing proton transfer and reducing internal resistance (Jang et al., 2004).

Figure 1.1 Schematic representation of an ML-MFC (a) single-chamber unit, (b) dualchamber unit. (Redrawn with modifications from Ref. (Kim et al., 2016)).

Figure 1.2 Advantages of MFCs. (Redrawn with modifications from Ref. W. W. Li et al., (2014)).

1.6. The process of bio-electricity generation

The process of bio-electricity generation in MFCs is similar, where the substrate with microorganisms is located in the anode compartment (if the anode is partitioned from the cathode using a separator) of the MFC chamber (Lin et al., 2013). Microorganisms play the role of a catalyst in the process of energy generation, producing electrons and protons along with carbon dioxide (CO_2) , and other inert gases. Over time, they create a biofilm on the anode surface, thus catalyzing the anaerobic bacterial oxidation (Z. Li et al., 2014; Waller and Trabold, 2013). The protons (hydrogen ions) $(H⁺)$ move from the anode towards the cathode, while the electrons transfer through an external electrical circuit connected to the cathode (Lin et al., 2013). Oxygen, which acts as the electron acceptor, is pumped from the air through the catholyte into the cathode compartment or is directly received if the cathode is in contact with the air (air-cathode) (Zhang and Ye, 2015). The basic chemical reactions taking place at the electrodes with the organic materials are depicted by Eq. (1.1). The specific reactions may vary in the anode compartment according to the type of substrate used; however, the products remain the same with altered stoichiometric coefficients (Khan et al., 2018). At the anode, the partition reaction in the case of glucose as substrate occurs as indicated by Eq. (1.1). The substrate, which functions as a fuel for the MFC unit, contains organic compounds that are degraded by the microorganisms during the process of

oxidation (Luo et al., 2007; Pauline and Boopathi, 2018). Agri-food industries produce a lot of waste easily broken down by bacteria and can be a good carbon source for MFCs. Other common substrates may include glucose, acetate, wastewaters from the domestic household, dairy, slaughterhouse, refinery, and dyeing industry (Jothinathan et al., 2018; Pallavi CK and Udayashankara TH, 2016; Savizi et al., 2012). In the cathode compartment, hydrogen ions and oxygen react with the help of electrons to form water or hydrogen peroxide (Eqs. 1.2 – 1.3) (Luo et al., 2017). The cathode electrode is generally coated with a catalyst such as platinum to assist in the reaction of hydrogen ions with oxygen or, in other words, to facilitate the oxygen reduction reaction (ORR) (Rahimnejad and Najafpour, 2011).

$$
C_6H_{12}O_6 + 6H_2O \rightarrow 6CO_2 + 24H^+ + 24e^-
$$
 Eq. 1.1

At the cathode, the partition reaction occurs as indicated by Eqs. 1.1-1.2.

$$
4H^{+} + 4e^{-} + O_{2} \rightarrow 2H_{2}O
$$
 Eq. 1.2

$$
4H^{+} + 4e^{-} + 2O_{2} \rightarrow 2H_{2}O_{2}
$$
 Eq. 1.3

Where, $\[\Delta G\]$ ^ $o = -5792.2 \; kJ/mol$

1.7. Influencing factors on MFC and ML-MFC

The performance of MFCs is influenced by a myriad of factors, encompassing chemical, physical, and biological elements that profoundly affect the overall efficiency of these units. Among the physical and chemical factors are reactor design (Jang et al., 2004; Liu and Logan, 2004; Luo et al., 2007; Ye et al., 2018), pH levels (Gil et al., 2003; Jadhav and Ghangrekar, 2009), temperature conditions (Feng et al., 2008; Jadhav and Ghangrekar, 2009) (Ahn and Logan, 2010), electrode surface area in conjunction with electrode spacing (Cheng and Logan, 2011; Ghangrekar and Shinde, 2007), pretreatment methods for influent (Ghangrekar and Shinde, 2008; Jadhav and Ghangrekar, 2009; Yang et al., 2013), as well as chemical treatments applied to the cathode and anode electrodes, and more.

Biological factors come into play as well, including the choice of microbial culture and the preparation of the substrate for these microorganisms (Hassan et al., 2014; Malvankar et al., 2012). Numerous studies have explored exogenous bacteria for their electron-producing capabilities (Cheng and Logan, 2007). A noteworthy challenge in ML-MFCs involves the mixing of the anolyte and catholyte, which can lead to a decrease

in cathode electrode efficiency due to biofouling. The primary objective of employing ML-MFCs is to mitigate internal resistance, optimize the conversion rate of organic materials within the substrate, reduce manufacturing costs, and enhance the efficiency of both anode and cathode electrodes (Rabaey and Verstraete, 2005).

Various methods exist for bacteria to transfer electrons to the anode, including electron transfer via a mediator, direct electron transfer, and electron transfer through bacterial nanowires (Khan et al., 2018). The thriving of bacterial cultures hinges on creating a conducive environment, with pH and temperature exerting significant influence on bacterial activity and, consequently, reactor performance. Multiple reviews have been published on MFCs, each emphasizing different aspects such as design features, substrates, microbial metabolism, performance, and associated challenges (Hindatu et al., 2017; Wei et al., 2011).

In this study, our focus centers on a physical factor that impacts the performance and efficiency of ML-MFCs, specifically the container or, more precisely, the wall of ML-MFCs. The use of wood as a container and separator has the potential to overcome cathode deterioration and the growth of microbial communities on the cathode while incorporating an air-cathode electrode.

1.8. Types of MFCs and ML-MFCs

Microbial Fuel Cells (MFCs) exhibit considerable diversity, and various classification schemes have been proposed in the literature. These classifications help us understand different aspects of MFCs, from their intended applications to their specific characteristics. In this section, we will explore classifications related to the presence or absence of membranes, the type of cathodes used, and their impact on bioelectric production and water treatment.

Figure 1.3 MFC classification according to membrane absence and air-cathode and aqueous-cathode design.

1.8.1. MFC with PEM/AEM membrane

One significant classification of MFCs involves the presence of Proton Exchange Membranes (PEM) or Anion Exchange Membranes (AEM). Numerous studies have delved into the comparison between MFCs with and without these membranes, revealing various advantages and disadvantages.

It is well-documented that membranes introduce increased internal resistance in MFCs (Liu and Logan, 2004; Tahir et al., 2022). Nevertheless, the benefits of using membranes in MFCs cannot be disregarded. These advantages include mitigating oxygen intrusion (Du et al., 2011), preventing cathode biofouling, and minimizing leakage issues (Yang et al., 2017).

Here, we will specifically discuss MFCs with membranes, categorizing them into two subtypes: air-cathode single-chamber and aqueous-cathode dual-chamber MFCs.

1.8.1.1. Air-cathode MFC

Researchers have explored various modifications of PEM/AEM membranes in air-cathode single-chamber MFCs. In one study, an MFC's cathode was equipped with a PEM (Nafion

117) membrane. This setup involved a single chamber with both the anode and cathode residing within it. The cathode, positioned between the substrate and the air, featured the Nafion membrane. The initial maximum voltage generation was lower compared to membrane-less units, but the MFC with the membrane exhibited a five-fold increase in voltage generation duration when compared to its membrane-less counterpart (Liu and Logan, 2004). This extended voltage output appeared to be influenced by glucose concentration.

Figure 1.4 Schematic diagram of an air-cathode MFC.

1.8.1.2. Aqueous-cathode MFC

In another study, an aqueous-cathode dual-chamber MFC was constructed, with an anodic and cathodic chamber separated by a cationic exchange membrane (CEM) (CMI-7000) to assess different cathode types. In this configuration, the anode and cathode occupy separate chambers, interconnected by a channel containing the CEM. Remarkably, this study demonstrated the successful treatment of untreated wastewater, achieving a maximum power density of over 27 W/m³ (Zhang et al., 2013). A schematic representation of an aqueous-cathode MFC with a PEM membrane is shown in Figure 1.5, with the cathode chamber positioned next to the anode chamber. Notably, air pumping into the cathode compartment is essential in all cases of aqueous-cathode MFCs.

It is worth noting that the majority of MFC studies focus on lab-scale prototypes due to the high cost of membranes. This fact makes Membrane-Less Microbial Fuel Cells (ML-MFCs) a more attractive option compared to traditional MFCs. To provide you with a clearer perspective, here are some examples of both lab-scale and pilot-scale experiments, as presented in Table 1.1.

Figure 1.5 Aqueous-cathode MFC with PEM membrane. Here, the cathode was placed next to the anode.

MFC	Anode	Cathod	Substrate	Volu	Water	Bioelectri	Referenc
design		e		me	treatment	$\mathbf c$	e
						productiv	
						ity	
Single	Graphit	$Cu-$	sodium	20 L	80% of	465	(Jiang et
chamber,	e rod	MnO ₂	acetate		contamina	mW/m^2	al., 2011)
air-		$Co-$			nt	500	
cathode,		MnO ₂			reduced,	mW/m^2	
ML							
Single	Carbon	Carbon	sodium	6.9 L		734 ± 54	(Thung)
Chamber,	Felt	plate	acetate			mV	et al.,
ML,							2019)
aqueous-							
cathode							

Table 1.1 Some pilot-scale and lab-scale of MFC and ML-MFC experiments.

1.8.2. ML-MFC of single/dual chamber with/without separator with affecting factors

In this comprehensive study, our primary focus has been the development of ML-Microbial Fuel Cells (ML-MFCs). The overarching goal of this research endeavor is twofold: firstly, to mitigate cathode biofouling, and secondly, to reduce the overall maintenance costs associated with ML-MFCs. The core of ML-MFC technology revolves around two pivotal aspects: electricity generation and water treatment capabilities. It is essential to strike a harmonious balance between these factors while ensuring the efficient production and transfer of electricity. Various influential factors come into play when optimizing the harnessing of energy and expediting the oxygen reduction reaction (ORR) in ML-MFCs. These factors include the choice of different substrates, substrate pretreatment methods, anode electrode performance, cathode ORR proficiency, and reactor design, among others. Notably, the design of the MFC itself is the most critical factor affecting electricity production, water treatment capacity, and the overall costefficiency of the process (Pant et al., 2010).

One of the most significant components that can be eliminated through thoughtful design considerations is the membrane. The presence of a membrane has a substantial impact on MFC productivity and costs. For instance, Song and his research team dedicated their efforts to enhancing cathodic ORR without the use of precious metals. They employed a carbon cathode as a replacement for platinum, given its exorbitant cost. In their comparative study, they examined a single-chamber MFC (SMFC) with an aircathode and no membrane against a two-chamber MFC (TMFC) equipped with a proton

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exchange membrane (PEM). The results were unequivocal, with the ML-MFC exhibiting the highest power output and ORR activity. The substantial internal resistance imposed by the ion-exchange membrane is evident, as reflected in the resistance values of 45 Ω and 80 Ω for SMFC and TMFC, respectively (Song et al., 2020). Another investigation demonstrated that an ML-MFC outperformed an MFC equipped with a ceramic separator and salt-bridge connection between chambers, yielding a 140% higher power output of 520–570 μW and greater stability (You et al., 2020).

Likewise, studies comparing MFC configurations employing a metal anode with a carbonaceous cathode have consistently favored ML-MFCs with membranes, primarily due to their superior performance in terms of maximum power density (MPD) (Yamashita et al., 2016). This underscores the direct correlation between the use of salt bridges, membranes, or separators and the escalation of internal resistance within the reactor—a detrimental factor hindering higher power output and overall efficiency (Min et al., 2005). As a result, it is not surprising that single-chamber ML-MFCs are now favored over their membrane-equipped counterparts, prompting recent research efforts to enhance their performance and facilitate their commercial viability (Logan, 2010).

As mentioned earlier, ML-MFC performance is influenced by an array of factors, and the final reactor output is a culmination of their synergistic interaction. This complexity makes direct comparisons of the individual factors across various studies challenging. Most research endeavors dedicated to ML-MFCs have systematically explored the effects of these diverse variables. Therefore, any evaluations and comparisons must be grounded in reasoned analysis, research outcomes, and established theories and facts. Now, let us delve into an exploration of the factors that have had a significant impact on ML-MFCs and the measures taken to expand their limitations and boost production efficiency.

1.8.2.1. Effect of Anode and Cathode Electrodes modification

Microbial Fuel Cells (MFCs) have emerged as a sustainable solution for electricity generation, offering a promising response to global environmental concerns (Palanisamy et al., 2019). Within the realm of MFC design, the choice of electrodes assumes a central role, posing both a critical challenge and an opportunity for achieving optimal performance. Efforts to enhance anodes have focused on fostering bacterial adhesion and enhancing the capacity for electron collection and transfer from both bacteria and the surrounding medium (Hindatu et al., 2017). On the other hand, for cathodes, performance in terms of the oxygen reduction reaction (ORR) significantly influences the coulombic efficiency (CE) of the MFC. It also plays a pivotal role in boosting the flow of electricity, ensuring durability, and maintaining long-term stability (Santoro et al., 2013) (Jiménez González et al., 2020).

The positioning of the cathode within the ML-MFC configuration is of paramount importance. Different designs come with their own sets of limitations. For instance, if the ML-MFC incorporates an air-cathode, it encounters challenges such as water loss, ORR activity, oxygen infiltration, cathode degradation, biofilm formation on the substratefacing side, and anode performance and bacterial culture support. Over the past two decades, numerous studies have diligently addressed these issues and proposed innovative solutions. Despite notable progress in addressing cathode performance issues, certain challenges persist, rendering MFC technology nonviable for large-scale applications (Zhou et al., 2012).

Researchers have explored various materials for developing electrode materials in MFCs, with carbonaceous materials (such as carbon cloth, carbon felt, and graphite) and metallic materials (including copper, zinc, and stainless steel) being among the most widely utilized (see Figure 1.6) (Santoro et al., 2017). Carbon, in particular, has garnered significant attention due to its abundance in nature and its exceptional properties, including its ability to interact with electroactive biofilms, high conductivity, and durability in harsh environments. As such, carbonaceous materials find extensive application as electrodes in MFCs (Santoro et al., 2017).

Various analytical techniques, including scanning electron microscopy (SEM), transmission electron microscopy (TEM), and X-ray diffraction (XRD), have been commonly employed to characterize electrode material structures. Most studies have evaluated electrodes' electrochemical performance, particularly in the context of ORR, using techniques such as cyclic voltammetry and electrochemical impedance spectroscopy. Some studies have expanded their focus to investigate oxygen diffusion through cathodes, MPD, CE, and chemical oxygen demand (COD).

Figure 1.6 Several carbonaceous and metallic materials are used for anode and cathode electrodes. (Published from Ref. Santoro et al., (2017) under CC BY 4.0 license).

1.8.2.1.1. Cathode Electrode

As mentioned, carbonaceous materials have unique properties such as electrical conductivity (Saba et al., 2017), corrosion resistivity (Slate et al., 2019), mechanical strength (Jia et al., 2018), high surface area (Yang et al., 2018), biocompatibility (Zhao et al., 2018), chemical stability (Cai et al., 2020b), environmental safety, and low cost (Liu et al., 2020). In a study, a carbon cathode was prepared from mixed carbon powder (Vulcan XC-72) and 30 wt% polytetrafluoroethylene (PTFE) solution, coated with different numbers of diffusion layers (DLs) from the airside. The optimum performance was obtained with four DLs, which significantly improved the Columbic Efficiency from 19.1% to 32%, while the MPD increased by 42% compared to an uncoated carbon cathode. Also, the open-circuit potential analyses revealed that the maximum potential difference between the cathode having four DLs and the uncoated cathode was 117 mV at 0.6 mA/cm2. The oxygen permeability and water loss from the cathode decreased with an increasing number of DLs (Cheng et al., 2006). Any attempt to seal the cathode further led to higher internal resistivity, like in the case of the MFC with PEM. A spun-bonded olefin sheet was used in

a study instead of a PTFE coating to seal the cathode to have lower resistivity. At the beginning (on day 5), this cathode produced an MPD of 750 mW/m², a current density of 2.0 A/m² (32 A/m³), a CE of 55%, low resistance of about 4 Ω , and total internal resistance of 103 Ω. However, the total internal resistance increased to 247 Ω after 50 days, while the current density was less than $1.0 \text{ A/m}^2 (16 \text{ A/m}^3)$ with a CE less than 35% (Oh et al., 2012).

Figure 1.7 Surface morphology of nitrogen- and phosphorus-doped ordered mesoporous carbon (NOPMC) illustrating the mechanism of the oxygen reduction reaction. (Reproduced with permission from Ref. Song et al., (2020) © 2020 Elsevier publisher).

With efforts to achieve high ORR activity at a reasonable cost. Recently, in a research conducted by Song et al., nitrogen- and phosphorus-doped mesoporous carbon (NPOMC) was used as a catalyst replacing platinum (Song et al., 2020). As elucidated in Fig. 1.7, the active phosphate sites on the mesoporous carbon surface are a doping centre for the ORR, where the nano porous structure of the catalyst helps intensify the ORR activity, and the nitrogen-doping sites cause a change in the charge distribution. The NPOMC in the ML-MFC gave only 30% of the ORR activity of Pt/C (154.0 mW/m^2) with 30–40% lower MPD. Moreover, the gradual formation of biofilm on the surface of the catalyst (biofouling) blocked the nano-structured active sites of the NPOMC and drastically affected its performance and stability. The electrochemical performance showed a current of 1.33 mA and internal resistance of 286 Ω after 30 days of operation (Song et al., 2020).

In another work, canvas cloth was used for the fabrication of electrodes. The nonconductive material was made to conduct electricity by coating it with one nickel or graphite and manganese dioxide $(MnO₂)$ as the catalyst. The nickel-coated canvas produced a better result than the graphite-coated canvas, with an MPD of 86.03 mW/m², a COD reduction of 95%, and a CE of 30.2% (Zhuang et al., 2009). Yet, in another study, bamboo charcoal coated with platinum was used as a cathode in an ML-MFC, giving maximum power and voltage of 1.16 mW and 0.50 V, respectively (Yang et al., 2009). Feng et al. employed stainless steel coated with polypyrrole/anthraquinone-2-sulfonate film as a cathode. It produced an MPD of 575 mW/m², and the cathode remarkably reduced oxygen and inhibited water leakage (Feng et al., 2011). Iron-based catalysts such as ricobendazole and niclosamide could be a possible alternative for platinum on the cathode electrode, as they have shown 20–25% higher efficiency than that platinum (Santoro et al., 2016).

Similarly, the ability of pristine graphene to enhance extracellular electron transfer has been exploited for platinum-free electrodes, producing a volumetric power of 3.51 $W/m³$ (Call et al., 2017). Thus, it can be concluded that sealing the air-cathode electrode, choice of electrode material, and cathode coating catalyst are crucial considerations for achieving a high ORR activity. Still, there is immense scope for improvement of the MFC with many undiscovered possibilities for the cathode, the main challenges being long lifetime, stable performance, and cost of the electrodes.

1.8.2.1.2. Anode Electrode

The anode material and its arrangement play a vital role in the attachment of the bacteria, enrichment of the biofilm, oxidation of the substrate, as well as the transfer of electrons between the bacteria and the electrode, which in turn affect the final output. Several methods of modification such as surface treatment, ammonia gas treatment (Cheng and Logan, 2007), heat treatment (Feng et al., 2010), acid treatment, electrochemical oxidation, surface modification with nanomaterials, and surface coating with conductive polymers (Hindatu et al., 2017) have been used with anode electrode to improve the overall performance of an ML-MFC. The anode electrode should be bio-compatible with the bacteria, in addition to the other primary factors used for evaluating the cathode electrode. (Kumar et al., 2013). Here, trying to discuss some research of how modifying the anode electrode impacted ML-MFC efficiency.

A study used an anode made of carbon cloth treated with ammonia, while the cathode was carbon cloth with a platinum catalyst. This combination substantially improved CE by 20% compared to the untreated anode and an MPD increment from 1640 to 1970 mW/m². The power attained was 7.5 times higher than the untreated electrode, and the start-up time was reduced by 50% (Cheng and Logan, 2007). In another study, a carbon fibre brush treated with acid and heat was used as the anode produced an MPD of 1370

mW/m² , which was greater by 34% and 7% than that achieved by using an untreated electrode and only a heat-treated electrode, respectively (Feng et al. 2010). Likewise, the anode was modified using nitric acid (CM-N) and ammonium nitrate (CM-A). CM-N performed better than CM-A giving an MPD of 792 mW/m² and CE of 24%, which were greater than that achieved with using an untreated anode by 43% and 71%, respectively. Also, the modified anodes showed deep cracks and rough surface that improved electron transfer (Zhou et al., 2012). An MPD of 1788 mW/ $m³$ was achieved in a similar study by treating the anode surface with nitric acid and ammonia (Yang et al., 2014).

Lin et al. worked as electrodes with different materials (stainless steel, copper, gold, and graphite carbon cloth). Copper anode showed erosion, while hindered electron transfer was observed in the case of stainless steel. On the other hand, the gold electrode showed very high performance. In terms of open-circuit voltage, the best result was obtained with a gold anode and a carbon cloth cathode, measuring in at 0.49 V (Lin et al., 2013). Another study used an anode of graphite felt coated with iron oxide ($Fe₂O₃$) and ferric oxyhydroxide (FeOOH) giving an MPD of 18 W/m³ (Wang et al., 2013). A voltage of 573 mV and an MPD of 884 mW/m² were obtained when polyaniline was used with a graphene-modified carbon cloth as an anode (Huang et al., 2016). A flame-oxidized stainless-steel anode resulted in an MPD of 1063 mW/m², which was 24% higher than the untreated anode. Furthermore, it was appreciably higher by 323% than the same MFC configuration with membrane (Yamashita et al., 2016).

Peng et al. added 5% of nickel-iron oxide (NiFe₂O₄) to the anode to achieve an MPD of 806.4 mW/m², reducing the internal resistance by 39% compared to the untreated anode (Peng et al., 2017). Using indium tin oxide coated glass as an anode produced the voltage and power output of 471 mV and 418.8 mW/m², respectively. However, its COD reduced efficiency was lower than the carbon brush (Jiang et al., 2018). Another study found that using carbonized cotton textile modified with molybdenum carbide nanoparticles as an anode delivered an MPD of 1.12 W/m^2 . This material offered a super performance in conductivity, high biocompatibility, electrochemical activity, and costeffectiveness (Zeng et al., 2018). Graphene coated with iron sulfide (FeS₂) nanoparticles was employed as an anode on different substrates, achieving an MPD of 3220 mW/m² and an outstanding current density of 3.06 A/m^2 with COD reduction of 1319 mg/l (Wang et al., 2018). Many studies have shown tremendous progress by improving the anode with different materials, nanoparticles and composites to give a superior electrochemical

performance. It is of prime significance to check the compatibility of the treated anode with bacteria, its conductivity and durability to ensure optimum efficiency.

1.8.2.2. Effect of operating temperature

Literature has shown that temperature greatly impacts the performance of ML-MFC by affecting the growth and survival of microbial communities, the conductivity of the substrate solution, internal resistance, and start-up time (Gadkari et al., 2020). A study on a single-chamber, air-cathode ML-MFC compared two operating temperatures (20 and 30 °C) and found a stable power density of 187 ± 8 mW/m² and CE of 10% at 30 °C. When the reactor was adjusted to 20 °C, MPD and CE decreased to 155 mW/m² and 8.9%, respectively. Moreover, as the temperature decreased from 30 to 20 ℃, the cathode potential dropped drastically (cathode potential MPD was reduced by 315% (from 20 to −43 mV, versus saturated calomel electrode (SCE) reference electrode)), while the anode electrode potential lessened by 21% (Feng et al., 2008). Another work examined ML-MFC under different temperature ranges, i.e., 20-35 and 8-22 ℃. The higher working temperature range resulted in higher COD reduction of 90%, lower current of 0.7 mA, and CE of 1.5%. On the contrary, at a lower temperature range, the COD reduced to 59%, the current rose to 1.4 mA, and CE increased to 5%. Thus, the lower temperature range favoured the current and coulombic efficiency, while higher temperatures worked better for COD reduction (Jadhav and Ghangrekar, 2009). However, a different trend was observed in a continuous flow ML-MFC operating at 30 ℃, which showed a power density of 422 mW/m² (12.8 W/m³), COD reduction of 26%, and CE of 1.7%. At ambient temperature (23 °C), the power density reduced to 345 mW/m² (10.5 W/m³), COD reduced to 19%, while CE dropped to a mere 0.7% (Ahn and Logan, 2010). Another study analysed the effects of varying temperatures in the range 15–35 ℃ and found the best results at 35 $°C$, showing MPD, current and CE of 74 mW/m³, 2.51 mA, and 10%, respectively (Tee et al., 2018).

The start-up temperature must be considered for maintaining longer power stability in batch and continuous mode ML-MFCs. Generally, the microbial community's sensitivity towards temperature decides the maximum operating temperature, beyond which the bacterial activity degenerates. In most studies, the observed optimum temperature ranges from 20 to 32 C favoured the growth of methanogenic bacteria and demonstrated a linear

rise in the power output. The highest operating temperature limit is usually 35 ℃, above which the reactor efficiency begins to decline.

1.8.2.3. Effect of pH

The bioactivity of the microbial community is considerably dependent on the pH of the substrate (Marashi and Kariminia, 2015). In a study, an increase of the pH value above 8.5 led to the precipitation of carbonates on the cathode surface in the batch mode of ML-MFC reactors. The deposition of the carbonate layer acted as a barrier and decreased the electrochemically active area, similar to the PTFE effect on cathode activity (Guerrini et al., 2015). Another work tested the effect of pH in the range of 5.5–7.5 and found that the internal resistance decreased as the pH difference between the anode and cathode solutions increased. The highest current was generated within an optimum pH range of 6.0–7.0, indicating lower microbial activity at sub-optimal pH compared to the optimal pH (Jadhav and Ghangrekar, 2009). Similar observations were found in yet another study that concluded the optimal pH of an ML-MFC as 7 (Gil et al., 2003). The effect of three different pH conditions (5.5, 7.0, and 8.5) was assessed on an ML-MFC. The highest power density was observed at pH 8.5, which was greater by 40% and 66% than the power densities observed at pH values of 7.0 and 5.4, respectively. This was evident as even though acidogenic and methanogenic bacteria are inactive in alkaline conditions, electrogenic bacteria are active (Marashi and Kariminia, 2015). When the pH of the anodic solution increased from 5.4 to 9.0 due to the hydrolysis of the urea, it caused a decrease in the anodic performance, implying that the anode can take a limited pH working range (Santoro et al., 2013). The acidogenic bacteria are active at pH 5.5, where hydrogen production dominates, overcoming the pollutants' degradation and decreasing COD removal, as compared to neutral or alkaline conditions (Marashi and Kariminia, 2015). In contrast, an alkaline environment is favorable for the electrogenic bacteria leading to a higher power generation (Yuan et al., 2011)

1.8.2.4. Effect of substrate pre-treatment

The quality of the substrate fed to the reactor is a crucial factor influencing the performance of the ML-MFC. The organic molecules need to be decomposed and dissolved in the substrate for the bacteria to be able to consume them. In a study, sludge was kept in the anode chamber for 15 days, followed by heat treatment at 100 ℃ and cooling to room temperature. It was then re-inoculated, while no inoculation was done for the cathode. As a result, the COD reduced efficiency reached 91.4% at an organic loading rate (OLR) of 2.65 kg COD/m³.d, giving a maximum power density of 6.73 mW/m² and a current density of 70.74 mA/m² (Ghangrekar and Shinde, 2008).

Similarly, a power density of 0.32 ± 0.01 W/m² was obtained with a sludge fermented for nine days at 30 ℃ before dilution with the primary effluent, whereas the untreated primary effluent gave a power density of 0.24 ± 0.03 W/m². The fermentation caused a reduction in the total suspended solids from 26.1 to 16.5 g/l and the pH from 5.7 to 4.5. Additionally, it increased the conductivity from 2.4 mS/cm to 4.7 mS/cm (Yang et al., 2013).

1.8.2.5. Effect of reactor design and different separators

The design of the ML-MFC reactor is the cornerstone for deciding the type of cathode material and assembly suitable for the reactor. The cathode can either be entirely submerged in the substrate or placed as an air-cathode electrode, where one half has contact with the substrate and the other half is exposed to the air. For the air-cathode, the ORR reaction occurs with the oxygen in the air; for the submerged cathode, the air is supplied through a compressor.

In a study, an ML-MFC was designed with a cylindrical shape having a diameter of 10 cm and a height of 100 cm. Graphite felt was used as the anode (surface area of 465 cm²) and placed at the bottom of the reactor; with glass wool (4 cm depth) and glass bead (4 cm depth) set above the anode, as shown in Fig. 1.8. The cathode, made of graphite felt (surface area of 89 cm²), was placed at the top of the reactor and the compartment was aerated. The electrode spacing was varied between 10 cm and 30 cm. The inlet of wastewater was from the bottom, and after passing through the layers, it was discharged from the top. The set-up yielded a power density of 1.3 mW/m² at a current density of 6-9 $mA/m²$ (Jang et al., 2004). A similar design by Ghangrekar and Shinde gave the maximum voltage of 358 mV, power density of 10.9 mW/m², COD reduction of 88%, and BOD reduction of 87% (Ghangrekar and Shinde, 2007). A two-chambered cylindrical ML-MFC of Plexiglas was designed with a diameter of 75 mm and a height of 100 mm. A carbon paper separated the two chambers, and the electrodes too were made of carbon paper. The cathode electrode was coated with platinum and dipped into the cathode chamber. After a 400 h long run, the maximum voltage output of 551 mV and power density of 121 mW/m²

were attained (Luo et al., 2007). In another work, the cathode and anode compartments were placed at different levels. The cathode chamber was located above the anode chamber so that the outlet of the anode was connected to the inlet of the cathode through a valve. The substrate was driven by gravity into the anode from the storage container placed at a higher level than the cathode. The air was pumped into the cathode chamber, and the influent entered from the anode and went through the connection to the cathode. The reactor assembly resulted in a maximum voltage of 160.7 mV, an MPD of 24.33 mW/m³, COD reduced efficiency of 90.45% (Du et al., 2011).

Figure 1.8 Cylindrical ML-MFC with anode and cathode placed inside the reactor. (Redrawn with modifications from Ref. Jang et al., (2004)).

The cylindrical reactors have the advantage of capturing the maximum number of protons escaping from the anode, where the influent enters from the bottom of the reactor and leaves from the top. The influent comes in contact with the cathode as it emerges out of the reactor, accelerating the movement of hydrogen from the anode to the cathode. Moreover, the speed of the fluid flow affects the electricity production and the treatment of influent. Despite many advantages of the cylindrical reactor, some drawbacks are associated with these designs. Firstly, the spacing between the two electrodes affects the reactor output and needs to be optimized. Secondly, there is a requirement for external air pumping into the cathode compartment. Thirdly, when the cathode is completely immersed in the substrate, it leads to the biofilm's rapid formation, resulting in biofouling, thus
lowering the performance of the cathode. Finally, the cathode area may not be enough for the reactor; the literature shows that the optimal ratio of the cathode size to the anode size must be 2:1 or higher (Cheng and Logan, 2011). Another variation in the reactor design could be wherein the cathode is placed outside the reactor. In a study, a cylinder-shaped reactor made of Plexiglass was designed having 3 cm diameter, 13 cm height, and evenly drilled holes on the wall, as shown in Fig. 1.9. The cathode was a cylinder wrapped with a flexible carbon cloth coated with C/Pt, as an air-cathode. Carbon granules were used as the anode with a surface area of 31 cm^2 . The inlet was from the bottom, while the effluent exited from the top. The reactor displayed a voltage of 0.384 V with a maximum volumetric power of 50.2 W/m³ at a current density of 216 A/m³ (with an internal resistance of 27 Ω) (You et al., 2007).

Figure 1.9 Cylindrical ML-MFC with anode inside and cathode placed outside the reactor. (Reproduced with permission from Ref. You et al., (2007) © 2007 Elsevier publisher).

Furthermore, this design helped to have a longer lifespan and higher cathode performance, simultaneously decreasing the overall internal resistance by minimizing the distance between the cathode and the anode. A modified ML-MFC configuration was developed to solve the issue of space between the cathode and anode. The design consisted of a twin cylindrical compartment with a volume of 1.85 l, an air-cathode made of carbon cloth coated with platinum from the air-facing side, and a brush-type anode made of carbon cloth connected with titanium wires. The anode and cathode were placed on each end of the compartment, with the cathode being at a distance of 1 cm from the anode. The design

resulted in an MPD of $39-53$ mW/m² from the cattle manure solid waste (Lee and Nirmalakhandan, 2011). Liu and Logan used a Plexiglass cylindrical container open from both sides with 4 cm length and 3 cm diameter, as shown in Fig. 1.10. The anode and cathode electrodes were placed on opposite sides of the cylinder. The anode was made of carbon paper, and the cathode was made of carbon cloth coated with platinum on the airfacing side. The inner side of the cathode was examined with and without PEM. The MFC, in the absence of PEM, could achieve an MPD of 146 mW/m^2 with 20% CE using domestic wastewater as the substrate; while in the presence of PEM, it produced an MPD of 28 $mW/m²$ at 28% CE (Liu and Logan, 2004). A similar reactor was built with a glass tube placed on top of the reactor (4 cm long and 1.4 cm inner diameter) with a perforated cap to help the aerobic bacteria access to air. The anode was made of carbon fiber brush, and the cathode was made of carbon cloth coated with platinum. The results showed an MPD of 268.5 mW/m² , COD reduction of 67%, phosphorus reduction of 97%, and ammonia reduction of 99% (Jiang, 2017). It implies that the productivity of small-sized reactors is generally higher than that of big-sized reactors due to the lower internal resistance of the former. However, a disadvantage of these designs is that mixing cathodic and anodic compartments is inevitable, thus creating biofouling on the inner side of the cathode.

Figure 1.10 Design of a lab-scale single-chamber ML-MFC. (Reproduced with permission from Ref. Liu and Logan, (2004) © 2004 ACS publisher).

In a different work, a micro-sized ML-MFC was designed with dimensions of $15 \times$ 5 mm, a thickness of 0.37 mm with a volume of 83 μl, and the anode and cathode were made of carbon cloth (Fig. 1.11). Glass fiber was placed between the electrodes to alleviate the mixing of the fluids from the two compartments and for assisting in hydrogen transfer. The electrons were transferred through the titanium foils connected to each electrode. The electrodes were held with the help of acrylic cover plates. The cell was sealed with a silicone pad stacked vertically, resulting in an MPD of 3.2 mW/cm^3 (Ye et al., 2018).

Many efforts have been made to utilize biochar as an electrode to develop economical and environmentally-friendly MFCs without compromising on their performance; parallel to this goal, a new design has emerged for MFCs. A study fabricated a cathode using a bamboo tube by carbonizing it at 900 ℃ in a nitrogen atmosphere followed by heat treatment at 350 ℃ to increase porosity (Fig. 1.12). The cathode was also brushed with polytetrafluoroethylene solution on the external side to make it waterproof. A carbon fiber brush was used as the anode. The cell produced an MPD of 40.4 W/m^3 and a CE of 55% (Yang et al., 2017). Biomass materials can be an excellent replacement for expensive electrode materials for ML-MFCs due to their renewability, wide availability, and low cost. One of the approaches to better design with a low price is preventing the mixing of anolyte with catholyte, minimizing oxygen intrusion, avoiding biofilm formation, and alleviating cathode deterioration. This can be achieved by putting a separator or an additional anode in the container to create a two-chamber ML-MFC (Kim et al., 2016). Recently, Nawaz and his team assembled a conical dual-chambered ML-MFC fabricated from graphite-based materials, as shown in Fig. 1.13. The cathode chamber was concentrically placed inside the anode chamber with 4 mm space between them. The anode was sealed from the top with an acrylic lid, and the air was pumped into the cathode. This design accomplished a treatment efficiency of 84.4% and MPD of 15.03 $W/m³$ (Nawaz et al., 2020).

Figure 1.11 Design of a millimetre scaled ML-MFC. (Published from Ref. Ye et al., (2018) under CC BY-NC 3.0 license).

Figure 1.12 (a) Fabrication process of the cathode from a bamboo tube; (b) MFC with a cathode made from the bamboo tube. (Published from Ref. Yang et al., (2017) under CC BY 3.0 license).

Figure 1.13 Schematic representation of an ML-MFC with two containers concentrically placed inside one another. (Reproduced with permission from Ref. Nawaz et al., (2020b) # 2020 Elsevier publisher).

1.8.2.6. Effect of electrode surface area and electrode spacing

The surface area of electrodes and the spacing between them are crucial factors to be taken into account for optimum performance of the ML-MFCs; it has to be chosen carefully where and in what anode/cathode ratio you will design the MFC container. A study investigated the effects of varying the electrode spacing (20, 24, and 28 cm) and anode area. It was found that more MPD could be achieved when the electrodes were placed close to each other. The maximum voltage of 358 mV was obtained at a distance of 20 cm. The power densities of 4.66, 6.45, and 10.13 mW/m² were achieved at anode surface areas of 210.64, 140.43, and 70.21 cm^{2,} respectively. It shows that the power density decreases with the increased surface area of the anode (Ghangrekar and Shinde, 2007). The trials for scaleup of ML-MFC showed that doubling the cathode surface area increased the power output by 62%, and doubling the anode surface area increased the power output by 12% (Cheng and Logan, 2011). In an experiment with a three-column MFC connected in series, an enhancement in the anode surface area from 360 to 1080 cm^2 in each column increased the maximum power output by 264% for column 1, 118% for column 2, and 151% for column 3. Also, the COD and BOD reduction efficiencies were increased by 137% for column 1, 279% for column 2, 182% for column 3, and 63% for column 1, 161% for column 2 and 159% for column 3, respectively (Gálvez et al., 2009).

1.8.2.7. Effect of Substrate Conductivity

The substrate's conductivity is another factor that profoundly influences the output performance of the ML-MFCs. One way to enhance the conductivity is by adding metal ions to the substrate. However, adding metal ions creates a toxic environment for the microorganisms. So MFCs are limited by the requirement to compromise between the toxicity issue and the conductivity of the substrate (Dong et al., 2015). It has been found that when the ions are added, the MPD increases linearly with rising electrolyte conductivity; the MPD escalated from 0.11 W/m^2 at a conductivity of 1e4 Sm1 to 1.02 W/m² at a conductivity of 1e2 Sm1. Further increase in conductivity beyond 2e1 Sm1 showed no more improvement in MPD (Gadkari et al., 2020).

In a trial, phosphate buffer was added to the substrate solution to increase its conductivity, which increased the power density from 1330 to 1640 mW/m² (Cheng and Logan, 2007). The MPD rose from 438 to 528 mW/m² when the phosphate buffer concentration was increased from 50 to 200 mM in a comparable experiment (Feng et al., 2008). The total power density was increased to 400 mW/m² by adding sodium acetate solution to the substrate (Jiang and Li, 2009). Likewise, putting as little as 0.5% of NaCl and Na₂SO₄ (1:1 ratio) changed the power density from 34 to 43 mW/m² and augmented the phenol removal rate by 4%. Increasing the amount of salt to 1% raised the MFC power density to 45 mW/m², whereas 2% salt resulted in an inhibitory effect on power generation (Du et al., 2015; Mousavi et al., 2016). Thus, inorganic salts can significantly improve the conductivity and decrease the resistance in the solution, thus boosting the efficiency of ML-MFCs.

Most research agreed on that one of the factors of power production from ML-MFCs is the result of increased conductivity, which in turn suggests greater ionic conduction and decreased ohmic loss. However, the substrate solution can only tolerate so much salt before it becomes unusable. As bacteria cannot tolerate high levels of salt in their environment, it is critical that the solution's high salinity does not negatively affect the bacterial colonies (Aaron et al., 2010; Mousavi et al., 2016). Therefore, it is important to carefully assess the bacterial response to external stimuli with respect to the compounds added to enhance the substrate's ionic strength.

1.9. ML-MFC applications in Water Treatment (Substrate)

The substrate plays a pivotal role in determining the efficacy of ML-MFCs in energy generation since it serves as the primary source of nutrients for the bacteria (Yang et al., 2009). ML-MFCs are designed to accommodate a wide range of substrates, which can consist of mixtures of various organic materials, wastewater, or lignocellulosic biomass (Pant et al., 2010). In contrast to traditional wastewater treatment methods, such as the energy-intensive aerobic-activated sludge process that necessitates significant air or oxygen supply to aeration tanks (Capodaglio and Olsson, 2020), ML-MFCs present a more sustainable approach. This work explores the potential of ML-MFCs in wastewater treatment by targeting the removal of organic compounds, inorganic substances, and heavy metals. Various researchers have evaluated ML-MFC performance using measurable parameters like current (or current density in $mA/cm²$ or $mA/m³$), voltage (open circuit or closed circuit), COD/BOD reduction, MPD, and CE, as well as inorganic or heavy metal reduction to assess substrate efficiency.

Numerous studies have focused on maximizing the efficiency of ML-MFCs for the treatment of wastewater containing organic matter while also generating electrical power. For instance, in a study, ML-MFCs fueled with glucose and wastewater achieved an MPD of 494 mW/m² and 146 mW/m², respectively, with corresponding CE values of 9–12% and 20%. The glucose-fed unit demonstrated an impressive glucose removal efficiency of 98%. Interestingly, when a proton exchange membrane (PEM) was employed in the same study, the MPD decreased to 262 mW/m² and 28 mW/m², with CE values of 40–55% and 28%, respectively (Liu and Logan, 2004). This observation suggests that the absence of a PEM led to an increase in MPD but a decrease in CE due to substantial oxygen diffusion into the anode. Furthermore, the glucose concentration was found to influence power output, with higher glucose concentrations resulting in improved MPD. Similar findings were reported in another study using acetate and butyrate as substrates, achieving MPDs of 506 mW/m² and 305 mW/m², respectively, which were 54% and 57% higher, respectively, compared to using a PEM (Liu et al., 2005). Feng et al. reported MPD values of 205 mW/m² for beer brewery wastewater, 494 mW/m2 for glucose (0.6 g/l), and 146 mW/m² for domestic wastewater, demonstrating an

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increased COD reduction efficiency from 54% to 98% as the strength of brewery wastewater increased from 84 to 1600 mg/l.

Utilizing coal-tar refinery wastewater as a substrate yielded an MPD of 4.5 $mW/m²$ at a voltage of 543 mV, along with an 88% COD reduction, 57% sulfate elimination, and 41% sulfur removal. Moreover, the ML-MFC efficiently removed over 90% of phenol and 2-methyl phenol (Park et al., 2012). When human urine was passed through an ML-MFC, it generated a current of 0.18–0.23 mA, reducing the initial COD of 10.9 g/l to 3.6 g/l after a 4-day retention period in the batch mode MFC unit. Another study investigating phenol concentrations in the range of 25–200 mg/l observed an increase in phenol removal from 80% to 97% as the phenol concentration increased from 25 to 100 mg/l. This setup achieved an MPD of 49.8 mW/m² and a current density of 292.8 mA/m² (Buitrón and Moreno-Andrade, 2014). Using rice straw pre-treated with acid (to degrade cellulose) as a substrate, an ML-MFC unit generated an MPD of 137.6 $mW/m²$ and achieved a COD reduction efficiency of 79% with an initial COD of 400 mg/l (Wang et al., 2014). Similarly, purified terephthalic acid wastewater produced an MPD of 65.6 mW/m² at a COD concentration of 8000 mg/l (Marashi and Kariminia, 2015).

Wastewaters from dairy, leather, and sewage industries, rich in organic materials, generated maximum power outputs of 1.98 mW, 1.95 mW, and 1.28 mW, respectively, accompanied by COD reduction rates of 85.4%, 80%, and 65% (T. Aswin et al., 2017). Other lignocellulosic materials used as substrates resulted in an MPD of 29 mW/m³ (Adekunle et al., 2016). When dye processing wastewater was tested at several organic loadings, it achieved an MPD of 515 mW/m², CE of 56%, and COD reduction of 85% at an organic loading of 1.0 g/l COD (Karuppiah et al., 2018). Additionally, a study involving petroleum refinery wastewater showed a treatment efficiency of 45.06% and a power density of 28.27 W/m³, while whey wastewater treatment resulted in 72.76% treatment efficiency and 23.23 W/m3 power density (Mohanakrishna et al., 2018). Wood, rich in organic matter, also displayed potential as a fuel for MFCs, with poplar wood achieving an MPD of 8555 mW/m² (Erensoy and Çek, 2018). Another investigation used

tomato waste as a substrate and produced an MPD of 60.041 mW/m², a current density of 99.174 mA/m², and a voltage of 0.701 V (Kamau et al., 2018).

Recently, several studies have focused on heavy metal (sulfur, copper, mercury) removal using ML-MFCs, given the extreme toxicity and carcinogenicity associated with these pollutants. Various heavy metals are released in the effluents from tanning, cement, electroplating, and dye industries. In one study, a substrate made from hydrolyzed heavy metal-containing wheat grain (HMWG) produced an MPD of 381 mW/m², CE of 15.7%, and COD reduction of 83.4% (Yuan et al., 2018). However, increasing the concentration of HMWG hydrolysate resulted in a slowdown of reactor electricity production. In another study, Cu (II) was used as an electron acceptor to investigate the mechanism of metal treatment in an ML-MFC. A low Cu (II) ratio achieved an 87.56% reduction in heavy metals, while an increasing Cu (II) ratio led to a drop in heavy metal reduction efficiency to 36.98%. This change also caused the voltage to decrease from 71 to 11.1 mV (Chan et al., 2020).

1.10. Research rationale & objectives

- \checkmark The discovery of various types of microbial fuel cells (MFCs) has opened up new avenues for applications in bioelectric harvesting, water treatment, hydrogen generation, desalination, and the removal of hazardous compounds from wastewater. In the construction of traditional ML-MFCs, plexiglass, glass, or plastic containers have typically been used in conjunction with different cathode and anode electrodes. However, these conventional ML-MFC designs utilizing these materials often suffer from cathode biofouling and cathode deterioration, which inevitably lead to reduced efficiency over time, rendering ML-MFCs commercially less viable.
- ✓
- \checkmark To address the limitations associated with cathode biofouling and deterioration, this research explores the utilization of wood as both a container material and separator, offering a sustainable and eco-friendly alternative. The primary objective of this study is to manufacture an ML-MFC using wood and to design the container in a manner that prevents direct contact between the cathode electrode and the substrate.

This innovative approach aims to investigate the potential utility of untreated wood as a solution for mitigating cathode deterioration and biofouling issues.

- \checkmark In light of these objectives, the following research goals have been formulated:
- \checkmark Preparation of wooden ML-MFC containers under varying experimental conditions.
- \checkmark Evaluation of the prototype's performance by assessing key response parameters, including open circuit voltage, current density, power density, and COD reduction.
- \checkmark Modification of the wooden ML-MFC by altering the wall thickness to optimize its performance.
- \checkmark Assessment of the wooden ML-MFC's capability for water treatment by measuring carbohydrate content in the substrate, including COD reduction, pH, conductivity, and sugar content.
- \checkmark Characterization of the wood plates before and after utilization using advanced analytical techniques such as SEM and DMA.

1.7. Dissertation outline

This dissertation has been structured into four chapters as follows:

i. Chapter I

This chapter sets the stage by introducing energy harvesting techniques and provides an overview of microbial fuel cells (MFCs) in general. It delves into the mechanisms of electric production and water treatment with MFCs and categorizes MFCs based on the presence or absence of membranes, air-cathodes, and aqueous-cathodes. Additionally, it briefly reviews various types of MFCs and focuses on ML-MFCs, highlighting the factors influencing electric production and water treatment quality. Finally, the chapter explores the diverse applications of ML-MFCs in wastewater treatment.

ii. Chapter II

This chapter details the experimental procedures employed for constructing wooden ML-MFCs and preparing the substrate. It provides a comprehensive description of the necessary tests conducted to evaluate the novel prototype's performance in terms of electric production and water treatment capabilities. Additionally, advanced analytical techniques are employed to characterize the wood plates before and after their use.

iii. Chapter III

In this chapter, the results and their subsequent discussion are presented. Key parameters related to electric production, water treatment capability, pH, conductivity, and water loss are analyzed and discussed, with a focus on the influence of different wood plate thicknesses in the fabrication of wooden ML-MFCs. Mechanical properties, EDX (Energy Dispersive X-ray), and SEM (Scanning Electron Microscopy) analyses of the wood plates are also included.

iv. Chapter IV

The final chapter presents the innovative outcomes achieved throughout the study. Additionally, recommendations for future research directions are provided, offering insights into potential avenues for further exploration and development in the field of wooden ML-MFCs and their applications.

CHAPTER II-MATERIALS & METHODS

2. Materials and Methods

2.1. Air-cathode WML-MFC fabrication

2.1.1. Preparation of electrodes

The anode electrode was prepared from carbon felt (PXFT-540, Zoltek Corporation, Bridgeton, USA), having a 114 cm^2 surface area -with 6 mm thickness- (assumed as a smooth surface). The cathode electrode was prepared from carbon cloth with a 162 cm² surface area. Both the electrodes were washed with deionized water and dried at 80 °C for 24 h. They soaked in 1M HCl for 24 h to remove the impurities on the surface of the electrodes. Finally, the electrodes were washed with deionized water and dried at 80 °C for 24 h. Four pieces of carbon felt as an anode electrode were connected with a copper wire and placed inside each WML-MFC container, as appears in Fig. 2.1. The anode occupies a total volume of 24 cm³ inside the containers. The cathode electrode was placed outside the container, covering it from all sides; as shown in Fig. 2.2, it was tightened to the container with a fabric thread. The cathode electrode is connected with a copper wire. In this design, no catalyst was loaded on the cathode, nor was any pre-treatment done on the anode and the wood plates (Fig. 2.3). From each WML-MFC container, the extended anode and cathode wires were connected to a data logger through a breadboard. Before the current and power measurements, different external resistors between $400 \Omega - 1 \text{ M}\Omega$ were tested to determine the optimal external resistor for WML-MFC.

Figure 2.1 The anode electrode preparation. (a) the carbon felt cut and immersed in HCl. (b) four sliced pieces of carbon felt joined through a copper wire.

Figure 2.2 Cross-sectional view of the WML-MFC designed in this study. (a) cathode electrode (b) wooden container (c) anode electrode (d) inlet/outlet for the substrate (e) copper wire extended from anode electrode.

2.1.2. Fabrication of the WML-MFC container using wood

When deciding the wood species to use in the fabrication of WML-MFCs, it is vital to consider factors such as cost, wood-water relation, and availability. For this study, the untreated wooden plates were obtained from three common tree species, viz., oak (*Quercus robur*), Scots pine (*Pinus silvestris*) and black locust (*Robinia pseudoacacia*). The wood plates were collected from different places which therefore the wood plates considered as to be from different trees, and all were the heartwood part of the tree. They were cut into plates with 2 mm, 3 mm, and 4 mm thicknesses with the required dimensions by a wood cutting machine and assembled with the help of silicone adhesive (Fig. 2.3). The container's dimensions were $37 \times 37 \times 59$ mm (± 1 mm). The inner volume of each container was 55 ± 1 5 ml. After placing the anode inside the container, the internal volume was reduced to $25 \pm$ 5 ml. Two holes were drilled on each box, one hole (9 mm diameter) for feeding of influent and discharging of effluent, and the other hole (2 mm diameter) for the copper wire extended from the anode electrode to pass through the container (Fig. 2.2). Three containers

of the same size were prepared from each of the three wooden species. The containers were used as such (with native wood) without any treated chemically or bleached, all used as their natural form. The experimental containers of WML-MFC have listed and annotated in Table 2.1.

Figure 2.3 Step-by-step preparation of WML-MFC. (a) Wood plates after they have been sliced using a wood cutting machine. (b) Wood plates glued together with silicone adhesive. (c) The anode electrode was inserted within the wood containers, and the sides were sealed. (d) The final appearance of the WML-MFC prototype after wrapping the WML-MFC with the carbon cloth cathode electrode.

Table 2.1 Description of all experimental containers.

2.2. Methodology for operating WML-MFC and analysis

2.2.1. Inoculation and operation of WML-MFC

The inoculation of each Wood-based Microbial Fuel Cell (WML-MFC) was carried out with a 20 ± 5 mL substrate blend. This blend was composed of 200 mL anaerobic digestion sludge sourced from a municipal wastewater treatment plant in Sopron, Hungary, 300 mL distilled water, outdoor mud, and sucrose as a carbon source. Prior to introduction into the WML-MFC, the substrate was allowed to settle for one hour to separate the mud. The experiments were performed in a batch mode with varying Hydraulic Retention Times (HRTs). Initially, each WML-MFC was loaded with 25 mL of substrate and operated for a week under open circuit (OC) conditions. This phase was critical for stabilizing the voltage and establishing a robust microbial community in the anode compartment, as reported by An et al. (2014). During the microbial growth phase, the substrate was refreshed every three days in WML-MFCs with 4 mm wall thickness. In contrast, WML-MFCs with 3 mm and 2 mm wall thickness required daily substrate renewal.

Upon transitioning to a closed circuit (CC), HRTs were adjusted between 20 to 120 hours. Each refill involved either replacing the effluent with a new substrate or using the remaining substrate from the previous day(s). Due to the high moisture diffusion through the wooden walls, daily substrate addition was necessary to maintain adequate submersion of the anode electrode. No chemical adjustments were made to influent pH or salinity. The oxygen required for the Oxygen Reduction Reaction (ORR) in these air-cathode ML-MFCs was drawn from the atmosphere. All experiments were conducted at ambient room temperatures, ranging from 15 to 25 degrees Celsius.

2.2.2. Data acquisition and analysis with portable data logger TDS-302 The TDS-302 portable data logger, equipped with a 10-channel switching box, was employed for measuring DC voltage and other parameters. The logger's operation was programmable to record data at specified intervals (Fig. 2.4).

During initial testing, voltage readings for each WML-MFC unit were manually recorded every 2 hours using a Tektronix MDO3104 oscilloscope. This process involved 15 containers representing three species, with five models each. Based on these preliminary results, bioelectrically productive containers were selected for further study. Continuous voltage monitoring was then conducted using the TDS-302 logger, programmed to record every 30 minutes. The optimal external load for power output was determined by varying the load from 400 Ω to 1 M Ω and monitoring the voltage stabilization for each load.

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The performance of the WML-MFCs was primarily evaluated based on their power output. The current (I) and power output (P) were calculated using Equations 2.1 and 2.2, respectively. Current density (j) and power density (Pd) were determined as a function of the anodic surface area (114 cm²) using Equations 2.3 and 2.4. The equations are.

$$
P = V \times I
$$
 Eq. 2.2

$$
j = V/(R \times A) \qquad \qquad \textbf{Eq. 2.3}
$$

$$
P_d = (I \times V)/A \qquad \qquad \textbf{Eq. 2.4}
$$

Here, V represents voltage (mV), R the external load (Ω) , and A the anode electrode surface area (m²).

Chemical oxygen demand (COD), pH, and conductivity of the influent and effluent were analyzed. Additionally, daily water loss through moisture diffusion from each container was recorded and used as a standard measurement.

2.2.3. Morphological analysis through scanning electron microscopy (SEM)

SEM, a method that scans materials with an electron beam to produce magnified images, was used for microanalysis and failure analysis of solid inorganic materials. In this study, SEM (S-4800, Hitachi) was employed to examine the inner wall of the wood containers and the cathode for bacterial growth and moisture diffusion effects on wood surface morphology and biofilm development on the cathode (Fig. 2.5).

Figure 2.5 Hitachi S-3400N scanning electron microscopy.

2.2.4. Viscoelastic behavior analysis through dynamic mechanical analysis (DMA)

Dynamic Mechanical Analysis (DMA) is a sophisticated technique that applies sinusoidal stress to a material and measures its phase lagging deformation, offering insights into the material's viscoelastic properties. This method, as detailed by Schlesing et al. (2004) and Sun et al. (2007), is instrumental in analyzing the physical characteristics of solids and polymer melts. It provides comprehensive data on modulus, damping, and can be tailored to measure various parameters such as force, stress, strain, frequency, and temperature effects. Significantly, DMA is invaluable for assessing the viscoelastic behavior of materials across a broad frequency spectrum.

In this study, the DMA (METRAVIB DMA 50), as shown in Figure 2.6, was employed to evaluate the average elastic modulus of oak and Scots pine wood samples, both pre- and post-experimentation. The DMA tests involved subjecting the wood samples to sinusoidal deformation at a constant temperature of 25 °C and a consistent moisture content of $12 \pm 2\%$. These tests were conducted over a frequency range of 1–100 Hz. The wood samples, with dimensions of 4 mm \times 45 mm \times 8.7 mm, were tested under a 3-point bending mode. For each frequency setting, ten readings of the storage modulus were recorded, and their average values were plotted against the respective frequencies, providing a comprehensive analysis of the wood's viscoelastic behavior.

Figure 2.6 Dynamic mechanical analysis machine of METRAVIB DMA 50.

2.2.5. Energy-dispersive X-ray spectroscopy (EDX)

Energy-Dispersive X-Ray Spectroscopy (EDX), often paired with electron microscopy, is a powerful analytical technique used for elemental analysis and chemical composition determination of samples. This method, which generates distinctive X-ray emissions to identify sample components, is widely integrated with scanning electron microscopy for enhanced analysis (Torres-Rivero et al., 2021).

In this research, EDX analysis was performed using the S-4800 Hitachi model (depicted in Figure 2.5) on the inner walls of the wood containers. The primary objective was to investigate the effects of the influent on the wood surface. To ensure accuracy and minimize potential errors, EDX analysis was conducted on at least two different sections of each wood sample. This approach was critical for identifying the chemical composition of the wood species used and measuring the elements absorbed by the wood after exposure to wastewater. This analysis provided valuable insights into the interaction between the wastewater influent and the wood material, contributing to a deeper understanding of the material's suitability and performance in the experimental setup.

CHAPTER III-RESULTS AND DISCUSSION

3. RESULTS AND DISCUSSION

3.1. Chapter Synopsis

This chapter provides a detailed analysis of the test results for Wood-based Microbial Fuel Cells (WML-MFC) using different wood types and thicknesses. We start by presenting preliminary findings, followed by a series of experiments focusing on Oak and Scots Pine wood at thicknesses of 2 mm, 3 mm, and 4 mm. These experiments aim to evaluate the effectiveness of using wood in WML-MFCs. Key performance indicators such as Open Circuit Voltage (OCV), Power Density (PD), Current Density (CD), and moistureinduced water loss are discussed, along with changes in Chemical Oxygen Demand (COD), pH, and conductivity in the influent and effluent. Additionally, the impact of wastewater on the wood and the cathode electrode surface is examined through Scanning Electron Microscopy/Energy Dispersive X-ray Analysis (SEM/EDX). Dynamic Mechanical Analysis (DMA) is employed to explore the viscoelastic properties of the wood samples.

3.2. Preliminary assessment of the 4 mm wall thickness WML-MFC

The feasibility of wooden microbial fuel cells was initially tested using 4 mm thick WML-MFC containers made from untreated European Oak, Scots Pine, and Black Locust wood. For each species, five sample containers were prepared. The Open Circuit Voltage (OCV) was the primary metric for assessing the concept's viability, recorded over 90 days at 2-hour intervals, manually with an oscilloscope. However, this section focuses on the final 210 hours of OCV data (refer to Figures 3.1, 3.2, and 3.3).

The preliminary tests showed that Black Locust exhibited significantly lower OCVs compared to Scots Pine and Oak. Specifically, Black Locust reached a maximum OCV of around 50 ± 20 mV upon fresh substrate addition, displaying brief stability before a rapid decline to as low as 1 mV at 22 ± 4 °C (Figure 3.1). In contrast, the maximum OCVs for Scots Pine and Oak were approximately 100 – 200% higher than those for Black Locust. These species also maintained electrical output stability 15 to 23 times longer after each substrate refresh (Figures 3.2 and 3.3). Oak showed greater stability, albeit with a lower maximum voltage than Scots Pine. Additionally, COD reduction efficiency was lowest in Black Locust. Consequently, due to its low electrical output, instability, natural bending when absorbing water, and minimal effect on COD reduction, Black Locust was excluded from further tests, focusing on Scots Pine and Oak.

Figure 3.1 Voltage of black locust containers during the last 210 h after 90 days of operating with open circuit condition.

Figure 3.2 Voltage of oak containers during the last 210 h after 90 days of operating with open circuit condition.

Figure 3.3 Voltage of Scots pine containers of last 210 h after 90 days of operating with open circuit condition.

3.3. The WML-MFC with different wall thickness

Initial trials with Oak and Scots Pine were successful. Determining a preference between Scots Pine and Oak required testing both species at multiple thicknesses. WML-MFCs were constructed with wall thicknesses of 2 mm, 3 mm, and 4 mm for both species to identify the optimal thickness. The experiment commenced with 4 mm thick WML-MFCs.

A consistent experimental setup was employed for all container variants to facilitate a fair comparison across different thicknesses. The OCV, CD, and PD were measured under batch mode operation, with Hydraulic Retention Times (HRTs) ranging from 24 to 72 hours. Containers were manually filled with substrate using a pipette. During each substrate cycle, the lid was opened, effluent was extracted, and new influent was introduced. All necessary precautions were observed during these substrate replacement processes.

3.3.1. Cell voltage with open-circuit and start-up progress to acclimation time 3.3.1.1. The WML-MFC with 4 mm wall thicknesses

In this section, we delve into the performance of Wood-based Microbial Fuel Cells (WML-MFC) with a specific focus on 4 mm wall thicknesses. The experiment, conducted under uniform temperature and influent conditions for all samples, evaluated Open Circuit Voltage (OCV), Current Density (CD), and Power Density (PD) across different wood species. It commenced with an open circuit, reaching a peak OCV after 351 hours,

albeit with some fluctuations. Among 4 mm thick prototypes, WP4-MFC-1 showcased the highest OCV at 0.551 V (see Fig. 3.4) and a notable COD reduction efficiency of 42%.

Subsequently, additional containers of oak and Scots pine with 4 mm thickness were connected to data loggers. Upon influent addition, a gradual voltage increase was observed, with notable surges upon each substrate replacement (Fig. 3.4), indicative of biofilm growth on the anode. WP4-MFC-1 attained an OCV of 551 mV after 280 hours, marked as the acclimation point. Comparatively, WP4-MFC-3 and WP4-MFC-2 reached 321 mV and 269 mV respectively, over longer durations. Among the oak containers, WO4-MFC-2 performed best, achieving an OCV of 141 mV within 343 hours.

This performance is noteworthy when compared with similar studies. For instance, Logroño et al. (2017) reported a maximum OCV of 577 mV using carbon felt electrodes and anaerobic sludge, while Wang and Su (2013) observed a maximum OCV of 102 - 246 mV in microfluidic ML-MFCs with different influents. A soil-based ML-MFC study by Simeon et al. (2019) yielded a maximum OCV of 729 mV under similar electrode conditions. These comparisons suggest that wood is a viable biomaterial for MFC or ML-MFC fabrication.

The difference in electrical production between oak and Scots pine was significant, with Scots pine generating 78-81% higher OCV than oak. This discrepancy could be due to two factors. Firstly, wood density varies among species, affecting hydrogen permeability (Sándor and Mihály, 2006). Secondly, the presence of antibacterial compounds in wood, which differ across species, impacts microbial growth (Prida et al., 2006; Sanz et al., 2011).

Voltage spikes were observed after effluent replacement, peaking within 25-30 hours before stabilizing. These fluctuations were quicker than reported in other studies (Nayak et al., 2018), possibly due to substrate conductivity, galvanic corrosion, oxygen intrusion, and moisture diffusion through the wood. Differences in anodic microorganism cultures and untreated electrodes may also contribute.

Voltage disparities among containers of the same wood species were attributed to the concentration of anti-bacterial compounds, wood anisotropy, and physical properties. Scots pine and oak's differing densities (330-520 kg/m³ and 697 kg/m³, respectively) influence their bioelectrical production (Sándor and Mihály, 2006).

Additionally, during biofilm development, voltage increases were noted, particularly in WP4-MFC-2 and WP4-MFC-3 (Fig. 3.4). These were influenced by daily temperature variations, as observed in previous studies (Ahn and Logan, 2010; Jadhav and Ghangrekar, 2009; Feng et al., 2008). Ambient temperature changes affect substrate temperature, impacting microbial activity, especially in softwoods like Scots pine.

Oxygen diffusion into the anodic chamber, a critical factor for microbial activity, differed between night and day (Liu and Logan, 2004; Fan et al., 2007; Lepage et al., 2012; Venkata Mohan et al., 2008). This variation was more pronounced in Scots pine due to its higher oxygen diffusion coefficient compared to oak (Gartner et al., 2004; Sorz and Hietz, 2006). At night, lower oxygen levels led to a slight voltage increase, which did not affect oak as much.

Overall, the findings emphasize the potential of using wood as an efficient and costeffective separator and container for enhancing cathode performance and durability in WML-MFCs, promising for sustainable bioelectricity generation.

Figure 3.4 Voltage during the start-up of 4 mm wall thickness oak and pine WML-MFCs with open circuit condition for 457 h until acclimation (red arrows indicate daytime, and purple arrows indicate nighttime).

3.3.1.2. WML-MFC with 2 mm and 3 mm wall thickness

This section focuses on the evaluation of Wood-based Microbial Fuel Cells (WML-MFC) with reduced wall thicknesses of 2 mm and 3 mm. The experiment aimed to understand how these variations in thickness influence the Open Circuit Voltage (OCV), Current Density (CD), and Power Density (PD) of the cells. Data were collected using a data logger, allowing for a detailed analysis of each container's voltage output potential.

The experimental procedure began in an open-circuit condition and was conducted over approximately 700 hours. Among the tested prototypes, WP3-MFC-2 and WP3- MFC-1 showed the highest OCVs at 0.147 V and 0.139 V, respectively (refer to Fig. 3.5). These values are notably lower than those observed for 4 mm thick WML-MFCs, indicating a potential correlation between wall thickness and voltage output.

A significant challenge encountered during data collection was the high level of noise (fluctuations) in the voltage recordings, potentially due to external electrical interference or data logger anomalies. Despite applying Data Analysis Exponential Smoothing to mitigate these fluctuations, clear night-and-day variations remained indiscernible. Nonetheless, a general trend was observed where WML-MFCs recorded slightly higher voltages during the day and at elevated room temperatures, with a slight decrease at night.

In comparing the performance based on wood thickness and type, it was noted that Scots pine with a 3 mm thickness generated higher voltages than its oak counterpart. Conversely, 2 mm thick oak containers outperformed those made from Scots pine. This observation suggests that both wood type and oxygen intrusion play crucial roles in voltage generation. Arranging the containers by wood density (from high to low) and maximum OCV (from low to high), a pattern emerges, indicating the significance of wood type in containers with a thickness of 3 mm or more. In contrast, for thicknesses less than 3 mm, wood density appears to be a more influential factor, affecting oxygen intrusion and water loss through diffusion.

Additionally, the ease with which hydrogen can reach the cathode layer is affected by the wood's density, particularly in thinner containers. This highlights the need for a cathode layer with high Oxygen Reduction Reaction (ORR) capabilities. Previous studies have shown that thicker polymer separators in MFCs reduce oxygen intrusion but are limited in power output due to increased resistance (Kondaveeti et al., 2017). Moreover,

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the development of microbial biofilms or fungi on the wood surface can impact oxygen penetration and wood degradation (Mohamed et al., 2021).

Intriguingly, with 2 mm thick Scots pine (WP2-MFC), a negative voltage was recorded (Fig. 3.8), a phenomenon not observed in 2 mm thick oak containers (WO2- MFC) (Fig. 3.7). This suggests that below a 3 mm thickness, excess oxygen intrusion interferes with the hydrogen in the anode chamber, causing ORR to occur within the container itself before hydrogen can reach the cathode. This observation opens up new avenues for exploring combinations of thin-wall Scots pine MFCs with membranes and cathodes optimized for high ORR, potentially enhancing the overall efficiency and performance of these innovative bio-electrical systems.

Figure 3.5 Voltage of 3 mm wall thickness Scots pine WML-MFCs with open circuit condition from the start-up until 700 h.

Figure 3.6 Voltage of 3 mm wall thickness Oak WML-MFCs with open circuit condition from the start-up until 700 h.

Figure 3.7 Voltage of 2 mm wall thickness Oak WML-MFCs with open circuit condition from the start-up until 500 h.

Figure 3.8 Voltage of 2 mm wall thickness Scots pine WML-MFCs with open circuit condition from the start-up until 500 h.

3.3.2. External resistor selection

The selection of the external resistor was crucial for optimizing the power output of the Wood-based Microbial Fuel Cells (WML-MFC). A range of resistors from 400Ω to 1 M Ω were tested to determine the most suitable one for each cell type. The testing involved WP4- MFC-1, WO3-MFC-2, WP3-MFC-1, WO2-MFC-2, and WP2-MFC-3, using a breadboard setup. For each resistor, voltage was recorded after a 5-10 minute interval to assess its impact on power density. Consequently, the Maximum Power Density (MPD) for each cell type was achieved with specific resistances: $330 \text{ k}\Omega$ for WP4-MFC, $51 \text{ k}\Omega$ for WO3-MFC, 20 kΩ for WP3-MFC, 820 kΩ for WO2-MFC, and 81 kΩ for WP2-MFC, as illustrated in Figure 3.9. Subsequent measurements of power and current were conducted using these optimal resistances for the respective cell types, revealing that the 3 mm wall thickness WML-MFCs demonstrated superior performance in terms of MPD.

Figure 3.9 Resistor selection assessment based on power density for; (a) WP4-MFC, (b) WO3-MFC, (c) WP3-MFC, (d) WO2-MFC, and (e) WP2-MFC.

3.3.3. Bioelectricity generation of WML-MFC as a closed circuit

3.3.3.1. The WML-MFC with 4 mm wall thickness

Upon switching from Open Circuit (OC) to Closed Circuit (CC) with selected resistors (330 k Ω for 4 mm wall thickness), WP4-MFC-3 exhibited the highest Maximum Power Density (MPD) and Maximum Current Density (MCD) of 35 mW/m² and 0.09 mA/m², respectively, after 397 hours. Comparatively, WO4-MFC-2 reached an MPD of 4 mW/m² and an MCD of 0.03 mA/m² after 404 hours (refer to Figures 3.10 and 3.11). The CC phase spanned 420 hours, involving five cycles of substrate replacement, necessitated by water loss through moisture diffusion.

The observed bioelectricity production and water treatment efficacy were modest, aligning with findings from previous studies which indicated limitations under low conductivity and untreated electrode conditions (Khater et al., 2022; Liu et al., 2022; Wang et al., 2022). Studies using plain carbon felt as electrodes reported lower voltage and MPD compared to treated carbon (Harshiny et al., 2017; Vicari et al., 2016). In contrast, a ceramic separator MFC achieved a higher MPD of 3.2 W/m^3 even in low conductivity conditions (Santoro et al., 2018), suggesting the potential of wood as a feasible material for MFCs, warranting further optimization.

The experiment revealed significant current fluctuations due to varied substrates and Hydraulic Retention Times (HRTs). Initially, fresh influent was used in all containers, with WP4-MFC-2 showing an initial high PD of 24 mW/m^2 , which then dropped to 2 mW/m^2 after 26 hours. Subsequent fillings with older substrates led to lower bioelectricity production. A notable cessation of current growth occurred when containers were fed with distilled water, indicating the washout of microorganisms from the anode. However, replenishing with freshly seeded substrate led to a gradual increase in CD, suggesting the re-establishment of biofilms on the anode.

The bioelectricity generation showed a marked increase in the final cycle, especially in WP4-MFC-3, which outperformed WP4-MFC-2 in the last two cycles. Oak containers, however, exhibited sluggish performance throughout. The experiment underscores the importance of inoculating the anode with a vibrant microbial culture and maintaining optimal HRTs for efficient bioelectricity generation. Ambient temperature fluctuations also impacted CD, as indicated by the rise in early morning (red arrow in Figure 3.11) and the decrease after sunset (purple arrow), emphasizing the sensitivity of WML-MFC performance to environmental conditions.

Figure 3.10 Power density of 4 mm wall thickness oak and Scots pine containers according to the surface area of the anode.

3.3.3.2. The WML-MFC with 2 mm and 3 mm thickness

This section examines the performance of Wood-based Microbial Fuel Cells (WML-MFC) with reduced wall thicknesses of 2 mm and 3 mm under Closed Circuit (CC) conditions. The experiment spanned 1000 hours, with non-uniform Hydraulic Retention Times (HRT) and daily refilling of containers to compensate for moisture diffusioninduced water loss. The resistors used were 820 kΩ for oak and 467 kΩ for Scots pine.

During this period, the highest Maximum Power Density (MPD) and Maximum Current Density (MCD) recorded for the 2 mm and 3 mm WML-MFCs were 5.14 mW/m^2 and 0.023 mA/m², and 4.7 mW/m² with 0.0029 mA/m², respectively (refer to Figures 3.12 to 3.17). Notably, the 3 mm Scots pine (WP3-MFC) outperformed the 3 mm oak (WO3-MFC) by approximately 500% in terms of current and power density. In contrast, the 2 mm oak (WO2-MFC) generated around 200% higher current and power density than the 2 mm Scots pine (WP2-MFC).

A unique observation was the occurrence of negative voltage in WP2-MFC, lasting over 300 hours (see Fig. 3.15). Additionally, the 2 mm WML-MFCs exhibited more significant voltage fluctuations with sharp peaks and rapid drops compared to the 3 mm models. These peaks and valleys were even more pronounced than in the 4 mm WML-MFCs, suggesting a correlation with wall thickness and material density.

The observed behavior indicates that lower-density container walls, such as those made from Scots pine, lead to increased oxygen intrusion, resulting in the Oxygen Reduction Reaction (ORR) occurring within the anode chamber. This phenomenon, coupled with enhanced hydrogen escape and greater water loss, underscores the challenges in using thinner-walled containers. Consequently, for containers with wall thicknesses less than 3 mm, it is advisable to use denser wood species or incorporate a Proton Exchange Membrane (PEM) into the system to mitigate these issues.

Figure 3.12 Current density of 3 mm wall thickness oak and Scots pine containers according to the surface area of the anode.

Figure 3.13 Power density of 3 mm wall thickness oak containers according to the surface area of the anode.

Figure 3.14 Power density of 2 mm wall thickness Scots pine containers according to the surface area of the anode.

Figure 3.16 Current density of 2 mm wall thickness oak containers according to the surface area of the anode.

These findings provide valuable insights into the impact of wall thickness and wood density on the performance of WML-MFCs. They highlight the need for careful selection of materials and structural design to optimize bioelectricity generation and overall system efficiency.

3.3.4. Effluent condition and wastewater treatment performance of WML-MFC

In this study, we assessed the wastewater treatment efficacy of Wood-based Microbial Fuel Cells (WML-MFC) by monitoring pH, conductivity, and Chemical Oxygen Demand (COD) of influent and effluent across different incubation periods in batch mode. Measurements were taken at various intervals after introducing new substrate and following each subsequent replacement.

Conductivity and pH Changes:

We observed a consistent increase in effluent conductivity across all container types. In the 4 mm wall thickness Scots pine containers, effluent conductivity nearly doubled that of the influent, while oak containers recorded increases between 27-100%. This rise can be attributed to influent concentration due to water loss through moisture diffusion and the release of wood extractives into the water. Similar trends were noted in 3 mm and 2 mm wall thickness containers, although the rise in conductivity was less pronounced.

Regarding pH, a divergence was noted based on the container wall thickness and wood type. In 4 mm oak and Scots pine containers, the pH tended to decrease. However, in 3 mm and 2 mm containers, we recorded an increase in pH from around 4 to approximately 5 ± 0.8 . This suggests the wood walls' substantial impact on normalizing the pH of the influent around 5 ± 0.4 , depending on its initial value.

COD Reduction Efficiency:

Focusing on 4 mm wall thickness WML-MFCs, which showed the best electrical performance, COD reduction was analyzed at different HRTs. In Scots pine containers, the COD reduction in effluent was 18% after 24 hours, increasing to 40% after 48 hours and reaching 48% in the 48-72 hour window. Conversely, oak containers demonstrated a slower COD reduction rate, with a 3% decrease after 24 hours, 22% after 48 hours, and 39% after 72 hours. These results indicate that Scots pine containers are more effective in water treatment within the first 48 hours.

However, after three days, the rate of water treatment slowed in all 4 mm containers. Contributing factors include the average daily water loss (Scots pine: 4.8 ml/day, Oak: 1.3 ml/day), leading to inactive anode regions due to air exposure, and galvanic corrosion of copper electrodes, which inhibits microbial growth. A continuousfed WML-MFC system could mitigate these issues by reducing oxygen intrusion, preventing copper oxide accumulation, and maintaining continuous substrate contact with the anode.

In thinner wall containers (2 mm and 3 mm), there was an unexpected increase in COD of the effluent. This contrast with the 4 mm containers suggests that wall thickness is a critical factor in WML-MFC design, especially for water loss management and treatment efficiency, with Scots pine showing particular promise.

This comprehensive analysis underscores the importance of wall thickness and wood type in designing WML-MFCs for effective wastewater treatment. The findings highlight the need for strategic system design, including continuous feeding mechanisms, to address challenges like water loss and electrode exposure, which significantly impact treatment efficiency.

The							
incubation		Influent	Effluent			Influent	Effluent
period (or		COD,	COD,	Influent	Effluent	conductivity,	conductivity,
HRT), day	containers	mg/1	mg/1	pH	pH	μ S/cm	μ S/cm
	WO4-						
	MFC		4091	6.3	4.99	372	706
	WP4-						
$\mathbf{1}$	MFC	4240	3475	6.4	5.41	654	1202
	WO4-						
	MFC		3812	4.87	4.6	680	866
	WP4-						
$\overline{2}$	MFC	4941	2929	4.93	4.01	890	1306
	WO4-						
	MFC		400	6.73	5.75	429	809
	WP4-						
3	MFC	664	340	6.88	5.92	688	1012

Table 3.1 Influent and effluent conditions under different HRTs of one day to four days with 4 mm wall thickness WML-MFCs.

	WO4-						
	MFC	$\qquad \qquad - -$	\cdots	5.43 5.2		460	662
	$WP4-$						
$\overline{4}$	MFC	$---$	$\qquad \qquad - -$		6.36 5.38	538	1080

Table 3.2 Influent and effluent conditions under different HRTs of one day to four days different with 3 mm wall thickness WML-MFCs.

The					
incubation				Influent	Effluent
period (or		Influent	Effluent	conductivity,	conductivity,
HRT), day	containers	pH	pH	μ S/cm	μ S/cm
	WO3-	4.67	5.16	790	2370
	MFC				
$\mathbf{1}$	WP3-MFC	4.67	5.16	790	2270
	WO3-	4.67	5.05	790	2240
	MFC				
$\overline{2}$	WP3-MFC	4.67	5.04	790	2330
	WO3-	4.81	5.45	1090	1790
	MFC				
3	WP3-MFC	4.81	5.52	1090	1880
	WO3-	4.67	5.2	790	2170
	MFC				
$\overline{4}$	WP3-MFC	4.67	5.07	790	2370

Table 3.3 Influent and effluent conditions under different HRTs from one day to four days with 2 mm wall thickness WML-MFCs.

3.3.5. Effect of HRT on the bioelectricity production

HRT influence was assessed under four different HRTs (1, 2, 3, and 4 days). WP4-MFC-1 displayed the highest OCV of 0.551 V with HRT of 2 days with a high COD substrate, while WP4-MFC-3 showed the highest bioelectricity production with MPD and MCD of 35 mW/m² and 0.09 mA/m², respectively, with HRT of 2 days. The voltage increased with longer HRTs probably due to the longer contact time between biofilm and organic matter; under a stable condition of the substrate with no movement in the anolyte helps the bacteria to grow and create a denser network of biofilm (Li et al., 2013). In contrast, the quick and dramatic drop in voltage observed with WP2-MFCs and WO2-MFCs necessitates an HRT of less than a day, or better yet, a continuous-feeding system.

Over the course of three days, HRT was tested using distilled water to determine the relationship between the water level in the anode chamber and electrical production (to exclude the organic loading rate (OLR) impact). A one-day HRT was shown to increase the bioelectrical productivity of 4 mm wall thickness Scots pine WML-MFCs by 50% (Fig. 3.10, 3.11). This may be attributable to a combination of factors, but it is mostly focused on the level of the water as de-ionized water is used, firstly, keeping the containers full of anolyte helps the anode to stay fully immersed under the substrate for a longer period; water-loss in 4 mm wall thickness Scots pine and oak containers was around 40% and 10%, respectively, within 48 h (as indicated in Table 3.4). A larger percentage of water loss was seen in some containers of WO3-MFCs, WP3-MFCs, WO2-MFCs, and WP2-MFCs, respectively (Tables 3.5, 3.6). Also, having an anolyte in the container helps the wooden

walls of the containers become saturated with water, which restrains oxygen from intruding through the wood plate (Sorz and Hietz, 2006). Thirdly, Galvanic corrosion of copper within the anolyte inhibited microbial development due to the toxicity of copper oxide toward living cells, continuous influent (or de-ionized water) adding helps to dilute the copper dioxide (Kumar et al., 2021; Naz et al., 2020; Trevors and Cotter, 1990).

The bioelectricity and water treatment of WML-MFC were less effective when the HRT was more than two days due to the low water level in each container and escalating the galvanic corrosion of copper wire. However, from other research, other prototypes of MFCs demonstrated higher bioelectric outcomes and better water treatment performance with four days of HRT and more (Li et al., 2013; Sharma and Li, 2010). Because of the significant impact of the container wall thickness on the water loss ratio; a higher container wall thickness leads to less water loss. Eventually, this directly impacts bioelectric output and water treatment capability. This work was conducted briefly; further research would be necessary to understand HRT's effect on WML-MFC better and solve the challenges.

The	Water loss due to moisture diffusion through the wooden wall,								
incubation	ml								
period									
(HRT) ,	WP4-	WP4-	WP4-	WO4-	WO4-	WO4-			
day	MFC-1	$MFC-2$	$MFC-3$	$MFC-1$	MFC-2	MFC-3			
$\mathbf{1}$	4.9	4.3	5	1.5	1	1.2			
$\overline{2}$	10.7	9.8	8.8	3.3	4.3	4			
3	15	14.4	12.9	5.9	5.8	5.6			

Table 3.4 Water loss in batch-mode fed for 4 mm wall thickness WML-MFC containers.

Table 3.5 Water loss in batch-mode fed for 3 mm wall thickness WML-MFC containers.

	Water loss due to moisture diffusion through the wooden wall,								
	ml								
The									
incubation	WP3-		WP3- WP3- WO3-		$WO3-$	WO3-			
period	MFC-1			MFC-2 MFC-3 MFC-1 MFC-2		MFC-3			

(HRT), day						
1	3	\mathfrak{Z}	4.5	2.2°	2.1	4.5
$\overline{2}$	9.2	9	7	4.5	3.2	6.8
3	13.6	16.2	13.2	6	4.6	8

Table 3.6 Water loss in batch-mode fed for 2 mm wall thickness WML-MFC containers.

3.3.6. Effect of sugar quantity and culture density

The anode chamber of the 4 mm wall thickness WML-MFCs was refilled with new substrate once they began operating with OC. Measurements typically involved five effluent replacement cycles and nine replenishment cycles. In each effluent replacement or refilling cycle, the anode chamber is filled with either newly seeded sludge substrate or substrate from prior days or distilled water. When the effluent was changed out for newly planted influent, the current and power rose due to the high sugar content and dense mixed culture. Substrates with less sugar produced lower current density and output power density than those produced by substrates with higher sugar contents (Fig. 3.11, 3.12). The CD curved short and dropped quickly after the containers were replenished with distilled water between 164 and 222 h because the culture density inside the anode was decreasing.

3.3.7. Cathode condition during the experiment

A primary goal of this research was to assess the durability of the cathode electrode against microbial attack. After three months of operation, visual inspections showed no biofilm or microbial degradation on the cathodes of WML-MFCs (see Figures 3.18, 3.19). SEM analysis further confirmed these findings. Both control (unused) and used cathodes from oak and Scots pine WML-MFCs displayed similar clean, smooth carbon fiber surfaces. Some debris and wetness-induced remnants were noted, but no significant microbial damage. These results indicate that using wood as a separator and container in WML-MFCs could offer a cost-effective solution for maintaining cathode integrity while producing bioelectricity.

Figure 3.18 SEM images of 4 mm wall thickness WML-MFC carbon cloths of the cathode electrode, (a) Control sample, (b) WO4-MFC, and (c) WP4-MFC. Digital photographs of the cathode electrode, (d) Control sample, (e) WO4-MFC, and (f) WP4- MFC.

Figure 3.19 SEM images of carbon cloths of the cathode electrode from, (a) WO2-MFC, (b) WP2-MFC, (c) WO3-MFC, and (d) WP3-MFC. Digital photographs of the cathode electrode from (e) WO2-MFC, (f) WP2-MFC, (g) WO3-MFC, and (h) WP3-MFC.

3.3.8. Mechanical properties and morphological structure of wood plates After three months of operation, the mechanical properties and morphological changes of the wooden walls in WML-MFCs were examined using Dynamic Mechanical Analysis (DMA) and Scanning Electron Microscopy (SEM). DMA revealed a significant mechanical loss in oak, while Scots pine exhibited only minor changes (Fig. 3.20). This difference may be attributed to variations in moisture diffusion and its effects on the mechanical properties of the wood. Chemical reactions occurring within the wood during moisture absorption and desorption also contribute to these changes.

SEM analysis showed that used oak samples developed surface cracks, whereas Scots pine maintained a smooth surface (Fig. 3.21, 3.22). These cracks could significantly affect the durability and strength of the wood, particularly in thinner walls. Furthermore, both oak and Scots pine showed debris accumulation on their surfaces, with oak absorbing more color from the wastewater, indicating the presence of copper oxide from the anode connection (Table 3.7). These findings suggest that Scots pine is more durable against microbial attack and the impacts of moisture diffusion. However, further long-term studies are needed to fully understand the effects on wood species durability.

Figure 3.20 Average elastic modulus versus frequency for 4 mm wall thickness oak and Scots pine at constant temperature (25 ℃) and humidity (14% RH). Blue dots are rawoak before use in the samples. The red dots are oak after three months of continuous use. Grey dots are raw-Scots pine before use in the samples. Yellow dots are Scots pine after 3 months of continuous use.

Figure 3.21 SEM micrographs of the inner surface of 4 mm wall thickness WML-MFC after three months of use from (a) oak (b) Scots pine. Digital photographs from the container wall show the inner surface cross-section and the side view of (c) oak (d) Scots pine.

Figure 3.22 SEM micrographs of the inner surface of WML-MFC after two months of use from (a) WO2-MFC, (b) WO3-MFC, (c) WP2-MFC and (d) WP3-MFC. Digital photographs from the container wall show the inner surface cross-section and the side view of (e) WO2-MFC (f) WO3-MFC, (g) WP2-MFC, and (h) WP3-MFC.

3.3.9. The EDX analysis of WML-MFCs

Energy-Dispersive X-ray (EDX) analysis was conducted to investigate the elemental composition of the inner surfaces of WML-MFC containers made from oak and Scots pine. This analysis is crucial for understanding how these materials interact with the contents of the WML-MFC, particularly in terms of metal absorption from wastewater.

3.3.9.1. Elemental Composition:

Both oak and Scots pine naturally contain Carbon (C) and Oxygen (O) in varying percentages, as highlighted in Figure 3.23 and detailed in Table 3.7. For instance, the 2 mm oak container showed 37.51% O and 27.83% C. Notably, 23.04% Copper (Cu) was detected, presumably absorbed from wastewater after the anode copper wire underwent galvanic corrosion. Traces of Fluorine, Aluminum, and Iron were also found, albeit in smaller quantities. Similar patterns were observed in oak containers of 3 mm and 4 mm wall thickness.

3.3.9.2. Nitrogen Presence and Copper Absorption:

A unique finding was the presence of Nitrogen (2.47% and 3.80%) in the 4 mm wall thickness containers of oak and Scots pine, respectively, which was absent in thinner containers. This might be attributable to the higher percentage of domestic wastewater used in these containers.

Copper absorption varied across different containers and wood types. Oak with 3 mm wall thickness exhibited higher Cu absorption than its 2 mm counterpart. However, the 4 mm oak container showed the least Cu absorption compared to the 2 mm and 3 mm variants. In contrast, Scots pine absorbed less Cu than oak, by $21 - 39\%$ compare to Oak containers. The highest Cu absorption in Scots pine was observed in the 3 mm container (18.41%) .

3.3.9.3. Other Chemical Elements:

Besides C and O, with Scots pine containers elements like Fluorine, Iron, Aluminium, Nitrogen, Mercury, and others wewre observed. The distinct elemental compositions of these containers indicate different interactions with the wastewater constituents.

There are no known studies to date that directly compare with these findings regarding the absorption of heavy metals from wastewater by oak and Scots pine of various wall thicknesses. However, these results underscore an additional advantage of WML-MFCs: the ability of wooden walls to absorb metals, including heavy metals, from wastewater. This absorption process may not only aid in wastewater treatment but also potentially inhibit the growth of fungal and bacterial colonies on the wooden walls, thereby enhancing the longevity and durability of these containers.

	WO ₂ -	WO3-	WO4-	WP2-	WP3-	WP4-
	MFC	MFC	MFC	MFC	MFC	MFC
Elements	$(wt.\%)$	$(wt.\%)$	$(wt.\%)$	$(wt.\%)$	$(wt.\%)$	$(wt.\%)$
Oxygen	37.51	32.39	40.46	40.37	37.59	36.17
Carbon	27.83	21.64	39.72	37.38	37.09	33.89
Copper	23.04	23.57	13.14	14.90	18.41	12.42
Fluorine	0.36			0.23	0.03	
Iron	6.93	8.67	4.61	2.08	1.50	3.56
Aluminum	0.24	0.07	0.08	0.20	0.58	1.22
Nitrogen		0.04	2.47			3.80
Mercury				0.06	0.30	
Others	4.08	13.65	1.52	4.78	4.50	8.94

Table 3.7 EDX of oak and Scots pine containers.

Figure 3.23 EDX spectra of WML-MFCs after 3 months of use; (a) WO2-MFC, (b) WP2-MFC, (c) WO3-MFC, (d) WP3-MFC, (e) WO4-MFC, (f) WP4-MFC.

CHAPTER IV

CONCLUSIONS & FURTHER STUDY

4. CONCLUSIONS

4.1. Chapter Synopsis

This chapter proposes the main conclusions from the present research work, followed by the directions for further research. The findings were derived from studying wooden microbial fuel containers, using three different species as container and separator, and constructing the containers with different wall thicknesses. The underlying basis of the research is to develop and optimize the ML-MFC by protecting the cathode from deterioration and biofilm formation with a low-cost, sustainable, and eco-friendly solution.

4.2. Research conclusions and achievements

The present work has attempted to address the issue of cathode deterioration and biofilm formation in ML-MFCs by simultaneously using wood as a separator and container. Untreated wooden containers, carbon felt, and carbon cloth was chosen to check the capability and viability of this novel work. It is noteworthy that no chemical pretreatment was done to adjust the experimental parameters like conductivity, and pH. The WML-MFC container turned out to be a very promising biomaterial for the fabrication of ML-MFCs. For further understanding and development of WML-MFC, the effect of different wall thicknesses was tested depending on the bioelectric and water treatment capability. Comparing the wall thickness of 4 mm, 3 mm, and 2 mm, it was found that the electric production and water treatment capability were the greatest with the 4 mm wall thickness.

The study effort has been condensed into the following bullet points, which outline the accomplishments and implications.

1. The WML-MFC prototype showed the potential to provide full safety of the cathode from any biofouling, thus preventing cathode deterioration (**CT-CP4**) **(CT-PA)**.

2. The cost of building ML-MFC and its maintenance decrease enormously, saving the cathode from deterioration builds hope for commercializing ML-MFCs in the future. Further, a wooden membrane-less microbial fuel cell system is a cost-effective, and environment-friendly approach (**CT-BC2)**.

3. With Scots pine 4 mm containers, the highest COD reduced percentage was 48%, whereas, in the oak containers, it was 39% within a hydraulic retention time of 48 h. For this experiment, three wood species with 4 mm wall thickness were assessed (black locust, oak, and Scots pine). Black locust resulted in the lowest bioelectricity compared to the other species with the lowest stability on the pick of its highest OCV. The MPD (anode-based)

and OCV could reach 35 mW/m², 551 mV, and 4 mW/m², 269 mV (with 330 k Ω resistor) for Scots pine and oak, respectively. For black locust, the maximum OCV reached 51 mV, which could be sustained for 2 h maximum (**CT-JP1-submitted)**.

4. Water loss occurs due to moisture diffusion in wood. The amount of water loss is a function of the wall thickness as well as the wood species. The water loss is more at less wall thickness of WML-MFCs. This can be an issue, therefore it is important to feed the substrate to the system with continuous-feeding instead of patch-feeding **(CT-JP1 submitted) (CT-CP4).**

5. Wood, being an insulation material, can protect microbial activity against low ambient temperatures during winter; since low temperatures drastically affect the vitality of the microbial culture. Thus, WML-MFC can help in maintaining the performance efficiency even in cold weather (**CT-BC2)**.

6. The bioelectricity and water treatment of WML-MFC were most effective at HRT of 2 days with 4 mm wall thickness **(CT-JP1-submitted)**.

7. With 2 mm and 3 mm wall thickness still possible to have a successful WML-MFC which could be used for different purposes. The highest MPD and MCD, obtained with 2 mm and 3 mm wall thickness of WML-MFCs, were 5.14 mW/m² at 0.023 mA/m², and 4.7 $mW/m²$ at 0.0029 mA/m², respectively during 1000 h of experimentation. Although these parameters (MPD and MCD) are less than those obtained for containers of 4 mm wall thickness, they manifest the viability of the idea **(CT-PA).**

8. For effective water treatment, 4 mm is the minimum wall thickness required to fabricate WML-MFC **(CT-PA)**.

9. The highest electric production is possible with 3 mm wall thickness of WML-MFC in case of utilizing the right resistor. The 3 mm wall thickness of WML-MFC yielded the maximum power density (without the water treatment) **(CT-PA) (CT-JP1-submitted)**.

10. Containers with a wall thickness of less than 3 mm must be fabricated from a dense wood species, or PEM must be included in the system. Compared to MWL-MFCs with a wall thickness of 3 mm Scots pine MWL-MFCs with a wall thickness of 2 mm exhibited negative voltage for more than 300 h, as well as highly sharp peaks and rapid decreasesin voltage. In addition, the peaks and valleys of 3 mm MWL-MFCs were sharper than those of 4 mm MWL-MFCs **(CT-PA) (CT-JP1-submitted)**.

11. The wooden walls of the containers have a massive impact on the pH and normalise the pH of the influent around 5 ± 0.4 ; if the utilized influent's pH is less than 5, it will rise to 5 ± 0.4 ; if the influent's pH is greater than 5, it will fall to 5 ± 0.4 (**CT-PA**) **(CT-JP1-submitted)**.

12. Wooden walls of WML-MFCs absorb metals and heavy metals from the effluent. This aids the water treatment process and improves the water's quality, and it also protects the wooden wall from rust and decay and provide a longer life for the WML-MFC containers **(CT-PA) (CT-CP4) (CT-JP1-submitted)**.

4.3. Research limitations

Through the experiment, several challenges were met, which can be improved to enhance the performance efficiency of WML-MFCs.

1. Batch-fed mode and manual filling of the containers resulted in high voltage fluctuations. Inside each WML-MFC, high air diffusion took place during the replacement of wastewater. It is possible to avoid this issue by building the container with four sides of plexiglass material and two sides only with wood.

2. Untreated wooden walls are vulnerable to oxygen intrusion, leading to the wastage of organic materials in the anolyte and shifting the dominant microbial culture. It would be interesting, to use PEM between the wood plate and cathode layer; this may help to reduce the oxygen intrusion. Moreover, with wood treatment, WML-MFC can become more compact and microbial-protective.

3. Untreated cathode led to low ORR. Free hydrogen may have escaped the cathode without going through the ORR process. The top and bottom parts of the WML-MFC container could be changed to plexiglass. This will help in building a continuous-fed WML-MFC and gain more control over the direction of hydrogen flow and obtain less leakage.

4. Securing the connections between the wooden plates of the container with a stronger adhesive. With some prototypes, leakage was observed, this comes to the lack of adhesive strength between the wood plates.

5. Wood is an anisotropic material. The profile of antibacterial compounds may change from one point to another, thus creating inherent differences between containers. The extraction of antibacterial compounds may improve the quality of the WML-MFC prototype.

6. Low conductivity of the substrate causes an increase in the internal resistance inside the WML-MFC containers and reduces bioelectric production of it.

7. The occurrence or increasing of copper oxide amount inside the substrate can be avoided by changing the copper wire to titanium, carbon rod or stainless steel or other microorganism-friendly metals for binding the carbon anode electrode.

4.4. Recommendations for further study

More research is required on the invented WML-MFC model for evaluating the continuousfed mode, anode/cathode volume ratio, wooden wall thickness effect, using PEM around the container to block oxygen intrusion, different cathode and anode configurations, differently treated anode and cathode materials, other wood species can be tested with various substrates using different biocatalysts, different treatments of the wooden wall, and combining wood with plexiglass or any other material in the structure.

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